

1 ACCUMULATION OF TRACE ELEMENTS IN FEATHERS OF THE KENTISH PLOVER *CHARADRIUS*
2 *ALEXANDRINUS*

3
4
5 MARCO PICONE^{1*}, FABIANA CORAMI², CARLO GAETAN¹, MARCO BASSO³, ALBERTO BATTISTON¹,
6 LUCIO PANZARIN⁴, ANNAMARIA VOLPI GHIRARDINI¹

7
8
9 ¹ Dipartimento di Scienze Ambientali, Informatica e Statistica, Università Ca' Foscari, Campus
10 Scientifico via Torino 155, I-30170 Mestre, Venezia, Italia

11 ² Istituto per la Dinamica dei Processi Ambientali, Consiglio Nazionale delle Ricerche, Via Torino
12 155, I-30170 Mestre, Venezia, Italy

13 ³ Via G.B. Verci 25/4, 35128 Padova, Italy

14 ⁴ Associazione Naturalistica Sandonatese, c/o Centro Didattico Naturalistico il Pendolino, via
15 Romanziol 130, 30020 Noventa di Piave, Venezia, Italy

16
17
18
19
20
21
22 * Corresponding author:

23 Marco Picone

24 Tel. +39 041 234 7742

25 Fax. +39 041 234 8584

26 email: marco.picone@unive.it

27 ORCID ID: orcid.org/0000-0003-0957-7247

28
29
30 Declarations of interest: none

31

32 **Abstract**

33 A non-invasive study of trace element accumulation in tail feathers of the Kentish plover
34 (*Charadrius alexandrinus*) was performed along the coastline of the northern littoral strip of the
35 Venice Lagoon, with the aim to verify whether contamination may be a factor affecting
36 conservation status of Kentish plover populations.

37 Body burdens in feathers of 11 trace elements including toxic metals/metalloids and essential
38 elements (As, Cd, Co, Cr, Cu, Hg, Ni, Pb, Se, V, Zn) were quantified by ICP-MS, then
39 concentrations were normalized to feather's age calculated using ptilochronology in order to obtain
40 daily deposition rates.

41 Mercury emerged as a major threat to the conservation of the species, since average feather
42 concentration was clearly above the adverse-effect threshold associated with impairment in the
43 reproductive success in a number of bird species. Also Cd and Se occurred at levels that may impact
44 on the conservation status of the studied species at local scale, even if to a lesser extent than Hg.

45 Gender-related differences in trace element accumulation emerged only for As, although for this
46 element the risks associated to environmental exposure seem to be negligible.

47
48 Key words – bioaccumulation, mercury, selenium, cadmium, water birds conservation,
49 ptilochronology,

50

51

52

53

54

55 **1 Introduction**

56 The Kentish plover *Charadrius alexandrinus* (hereafter KP) is a small wader of Family
57 Charadriidae, breeding in wetlands and coastal areas of Europe, North-Africa, Middle-East and
58 Central-Asia (Delany et al., 2009). Population size of KP is declining all over its distribution range,
59 and the cause of the decline has been recognized in habitat loss and fragmentation, increased human
60 uses of the sandy coastal areas for commercial and recreational purposes, and increased predation
61 by birds and mammals taking advantage of human activities (Delany et al., 2009; Domínguez and
62 Vidal, 2003; Scarton, 2017). For this reasons, the species **was** listed in the Annex I of the Directive
63 2009/147/EC and **was** included in the IUCN Italian Red List as an endangered (EN) species
64 (Rondinini et al., 2013).

65 Biology, breeding ecology and behaviour of KP have been widely studied in all its distribution
66 range (Argüelles-Ticó et al., 2016; Fraga and Amat, 1996; Kosztolányi et al., 2009; Lessells, 1984).
67 Nevertheless, no researches **were** performed up to date to assess the impact of environmental
68 contamination on its conservation status, and few data are available concerning bioaccumulation
69 (Kim and Koo, 2008; Zheng et al., 2018).

70 In coastal areas and wetlands KP feeds mostly on polychaetes (nereidae), small crustaceans
71 (isopods, ostracods and amphipods), small bivalves, coleopterans and dipterans (Castro et al., 2009;
72 Cramp and Simmons, 1983; Perez-Hurtado et al., 1997). Although KP is not a top predator in
73 **coastal** food webs, its dietary habits make it susceptible to bioaccumulation and, possibly,
74 biomagnification of toxicants. In estuarine and coastal areas, trace elements are of concern due to
75 their ubiquitous distribution and the potential detrimental effects they can exert on birds physiology
76 (Furness, 1996; Ohlendorf and Heinz, 2011). Toxic effects from trace elements includes changes in
77 behaviour, impairments in development and reproduction. Mercury (Hg) has been proven to cause
78 embryo malformations, lowered hatchability, decreased chick growth, reduced survival of young,
79 changes in reproductive behavior and endocrine disruption in waterbirds (Burger and Gochfeld,

80 1997; Jayasena et al., 2011). Lead (Pb) may disrupt the heme biosynthetic pathway and cause also
81 death (Blus et al., 1995 and citation herein). Cadmium (Cd) may induce growth retardation,
82 anaemia, suppression of egg production, kidney damage and marrow hyperplasia (Furness, 1996;
83 Spahn and Sherry, 1999). Selenium (Se) is highly toxic for the early developmental and embryonic
84 stages and causes reproductive impairments, reduction of hatching success and teratogenicity
85 (Ohlendorf and Heinz, 2011; Spallholz and Hoffman, 2002). For many other elements little is
86 known about possible detrimental effects in birds.

87 Trace elements concentration in birds may be measured in various organs (liver, kidneys, brain),
88 tissues (muscle, bones, fat), excrements, feathers and eggs (Dauwe et al., 2000; Markowski et al.,
89 2013; Zheng et al., 2018).

90 Feather analysis is a non-destructive and ethically preferable techniques as compared with organ
91 and tissue analysis, since feathers may be easily collected and, if necessary, repeatedly sampled to
92 study the accumulation of trace elements without affecting welfare or survival of the individuals
93 (Adout et al., 2007). A feather is connected with the blood circulation only during its growth period;
94 in this phase elements taken up through the food are incorporated into the sulphhydryl rich
95 keratinous matrix of the developing feather, where they are stored. When the feather has completely
96 grown, the vascular connection shrivels up and trace elements cannot be further allocated into the
97 keratin (Burger, 1993). Thus, their accumulation in feathers may be associated with a definite time-
98 span, corresponding to the time needed to fully develop the feather, variable from species to species
99 (García-Fernández et al., 2013; Grubb, 2006). This time-span can be determined using
100 ptilochronology, the study of the alternating pale and dark bands of feathers which is indicative of
101 the daily growth (Grubb, 2006), so that the deposition rates of the contaminants can be calculated
102 (García-Fernández et al., 2013), and the dilution artefact due to feather growth rates and mass
103 reported by Bortolotti (2010) can be minimized. The author, indeed, postulated that for elements
104 and compounds functionally incorporated in the feather (as sulphur and other elements), the

105 deposition should follow a *mass-dependent* model, whilst the deposition of non-structural elements,
106 as contaminants and hormones, occurs incidentally and should be *time-dependent*, so that growth
107 rate of the feather becomes a critical parameter (Bortolotti, 2010). Moreover, in the case of resident
108 populations and non-migratory species, the determination of trace elements in feathers represents a
109 reliable method for monitoring the uptake of contaminants through the local food web (Rothschild
110 and Duffy, 2005).

111 To verify whether KPs breeding in the area of the Venice Lagoon are exposed to levels of trace
112 elements that may, directly or indirectly, contribute to the local decline of the species - estimated by
113 Scarton (2017) as loss of 30% of the breeding pairs in the last three decades - an accumulation study
114 using feathers collected from specimens breeding along the littoral strips of the lagoon was
115 performed.

116 The focus was mainly on those elements known to cause detrimental effects on birds, for which
117 threshold concentrations for adverse effects are available in the literature (such as Hg, Cd, Pb and
118 Se); the concentrations of metals and metalloids possibly involved in toxic effects, but for which no
119 threshold data are available (including As, Co, Cr, Cu, Ni, V, Zn) were measured as well.

120 The extent to which KP deposited contaminants in the feathers was also assessed by comparing
121 trace element concentrations in KP feathers with literature data concerning other waterbirds in
122 different geographical areas.

123 **2 Materials and methods**

124 2.1 Study area

125 Feathers have been collected in the Cavallino-Treporti peninsula, the northernmost littoral strip
126 separating the Lagoon of Venice from Adriatic Sea (Figure 1). Along the 13 km of the coasts of the
127 peninsula, develops a fragmented system of natural and restored dunes covering a surface of about
128 350,000 m² (Cecconi & Nascimbeni 1997). Here the sandy substrate is characterized by the typical

129 herbaceous vegetation of drift lines (habitat "1210 - Annual vegetation of drift lines" according to
130 the Directive 92/43/EEC), embryo and shifting dunes (habitat "2110 - Embryonic shifting dunes"
131 and "2120 - Shifting dunes along the shoreline with *Ammophila arenaria* (white dunes)"). The
132 semifixed dune system is occupied by the habitat "2230 - Malcolmietalia dune grasslands" and by
133 the occurrence of the endemic, priority habitat "2130* - Fixed coastal dunes with herbaceous
134 vegetation (grey dunes)" (Del Vecchio et al., 2015).

135 In this area a small population of KP, estimated in 13-16 pairs in the period 2016-2018, nests
136 regularly on breeding grounds overlapping drift lines, embryonic and shifting dunes, and partially
137 also semifixed dunes. KPs feed all over the year on invertebrate assemblages colonizing shallows of
138 the northern basin of the Lagoon (mostly infaunal polychaetes and crustaceans), but also on
139 invertebrate communities of shores (infaunal polychaetes, crustaceans, and occasionally insects),
140 and on drift materials deposited by the sea.

141 Trace elements concentrations in shallow sediments within the Lagoon are well known for As, Cd,
142 Cu, Hg, Pb and V, whilst data are deficient (or lacking) for many others, such as Se. Shallows and
143 mudflats of the northern basin are generally characterized by low concentrations of metals as
144 compared with the industrial district and urban areas (Picone et al., 2018), with only As and Hg
145 occurring at concentrations above effect range low (ERL) and effect range median (ERM) (Long et
146 al., 2006). Nevertheless, recent studies evidenced that trace elements may biomagnify (Dominik et
147 al., 2014) and exert toxic effects toward aquatic invertebrates (Picone et al., 2016, 2018) also in
148 shallows far from the urban centres and industrial districts, and used by birds for feeding. In
149 contrast, very few data are available for sediments of littoral strips, and in most cases the
150 information refers to old studies (Donazzolo et al., 1981).

151 2.2 Sampling

152 Fully grown feathers of KP were collected during the breeding season, from March to September
153 2018. Adult birds were trapped at the nest during incubation of clutches, females in the morning and
154 males at dusk. Trapped bird were sexed, weighted at the nearest gram and measured for the standard
155 determination of head, bill, wing and tarsus length. A metal ring and a yellow colour ring with a
156 black, 3-digit alphanumeric code were also applied on the left and right tarsus, respectively, to
157 univocally identify birds and avoid duplicate sampling during the same season. A total of 13
158 individuals were captured (8 females and 5 males) representing the 48% of the total population
159 nesting in the area during the 2018 breeding season. This sample size is also very similar to that
160 used by Lucia et al. (2012, 2014) for studying accumulation patterns in other Charadriidae species.

161 The right outermost feather of the tail was removed from each bird. Tail feathers were chosen since
162 growth bars are more evident in the outermost rectrices (Grubb, 1989). Once collected, feathers
163 were individually placed in paper envelopes labelled with date, ring codes and sampling location,
164 and stored at room temperature until growth rate determination and chemical analyses (Jaspers et
165 al., 2006).

166 Removal of the feather caused no visible trauma to birds and did not affect their flight ability.

167 2.3 Ptilochronology

168 Growth bands were identified and measured following the procedure reported in Grubb (1989,
169 2006). Each feather was taped by its calamus to a millimetre graph paper sheet. An insect-mounting
170 pin was perpendicularly pushed at the proximal and distal ends of each feather to mark on the
171 millimetre graph paper its total length. Then, the pin was also pushed through the feather at the
172 distal edge of each dark bar, to mark on the paper the distance between the consecutive growth bars.
173 According to Grubb (1989), the point at 2/3 of the feather's length from the proximal end was also
174 marked. The width of growth bars was determined for a minimum of 5 to a maximum of 10 growth

175 bars centred at the 2/3 point, using a software for digital images analysis (Motic Images Plus 2.0).
176 The total length of the feather **was** measured to the nearest 0.01 mm, using a digital caliper. The
177 average width of the growth bars was calculated for each feather; the time needed to fully develop
178 each feather was then calculated by dividing total length of the feather by average width of growth
179 bars.

180 2.4 Chemical analyses

181 KP feathers needed to be completely mineralized and then analysed by ICP-MS (Agilent 7500Ce),
182 in order to evaluate body burdens of the trace elements of interest. Microwave digestion (Ethos 1,
183 Milestone) is widely acknowledged as the most robust sample preparatory technique for ICP-MS.

184 Every feather was weighed as a whole and then decontaminated before microwave digestion; all the
185 instruments, watch glasses and Petri capsules were previously decontaminated, rinsed with
186 ultrapure water (Elga) and dried prior to be used. In order to remove any impurity, every specimen
187 was thoroughly washed, alternating ultrapure water (Elga) and a solution of acetone 58.08 g l⁻¹
188 (1M) (Acetone 99.9 SpS Romil); to remove any residual trace of acetone, every feather was lastly
189 rinsed with ultrapure water and put inside the mineralization vessels.

190 For the mineralization of feather, the vessel-inside-vessel technologyTM was employed (Milestone,
191 Ethos 1); the secondary quartz vessels contained the feather and digestion reagents (Nitric Acid
192 HNO₃ 69% Plasma Pure Plus, SCP Science, and Hydrogen Peroxide H₂O₂ 30-32% Ultrapure
193 Romil, ratio 4:1), while the primary TFMTM vessel contained the solution required to accomplish
194 accurate temperature monitoring (Ultrapure water, Elga, and H₂O₂ Ultrapure Romil, ratio 4,5:1).
195 Reagent blanks were similarly prepared, excluding the sample, and were placed in the vessels in a
196 random position, to test any matrix effects. Microwave digestions were performed at controlled
197 temperature with a high pressure rotor. Once the mineralization was done and cooled, every

198 digested sample was recovered, weighed, suitably diluted (with ultrapure water, in triplicate) and
199 stored frozen at -20°C till the analysis in ICP-MS.

200 Primary and secondary vessels were cleaned after every digestion; after being thoroughly rinsed
201 with ultrapure water, quartz vessels were filled with ultrapure HNO₃ and primary vessels were filled
202 with the above described mix. An appropriate digestion procedure was then performed. Then all
203 vessels were thoroughly rinsed with ultrapure water and let to dry, prior to the next digestion. All
204 the steps were performed under a cleaned fume hood, in order to minimize any contamination.

205 A Certified Reference Material, ERM-CE278k (Joint Research Centre Institute for Reference
206 Materials and Measurements, Geel, Belgium), was analyzed to check for the accuracy of the
207 procedure. Results obtained showed good recoveries for all trace elements analysed in feathers (>
208 85%), thus the procedure was considered as accurate.

209 Measurement error, expressed as the percentage of the relative standard deviation (σ^* , RSD) was
210 <10% for all elements of interest; measurement uncertainty, expressed as the percentage of the
211 standard error (SEM) was < 5% for all elements studied and the confidence interval was $\geq 95\%$.
212 According to the results obtained, the method resulted repeatable and reliable.

213 Daily deposition rate (DDR) was calculated according to the following equation:

$$DDR = \frac{\text{trace element concentration in feather (ng g}^{-1}\text{)} \cdot \text{feather mass (g)}}{\text{feather's age (d)}}$$

214 2.5 Data analysis

215 One-way analysis of variance (ANOVA) was used to compare the average width of growth bars
216 among individuals of KP; prior to ANOVA, normality and variance homoscedasticity of raw data
217 were verified using Kolmogorov-Smirnov and Bartlett's test, respectively. Non-parametric Mann-
218 Whitney U-test ($\alpha = 0.05$) for independent samples comparison was used to compare the average
219 concentration of trace elements in KP's feathers between sexes. Spearman's non-parametric

220 correlation was used to assess possible correlations among trace elements in feathers of KP, and
221 between trace elements and weight. *Post-hoc* power analysis was used to assess power variability as
222 a function of sample size ($n = 13$), difference between males and females trace element
223 concentrations and standard deviation.

224 **3 Results**

225 3.1 Ptilochronology

226 The average length of the sixth rectrix of KP was 51.6 ± 1.6 mm ($n = 13$). Nevertheless, growth
227 bars were clearly discernible only in six rectrices (4 retrieved from males, 2 from females); since no
228 significant differences have been observed among individuals (one-way ANOVA: $F_{5,32} = 0.729$, $p =$
229 0.232), the pooled average width of 2.38 ± 0.30 mm was used as daily growth rate for estimating
230 feather age. According to these data, the average feather age for rectrices of KP has been estimated
231 to be 21.7 d (min. = 20.6 d; max. = 22.9 d).

232 3.2 Trace elements in feathers of Kentish plover

233 All analysed trace elements occurred at detectable concentrations in tail feathers of KP (Table 1).
234 As expected, trace elements deposited at higher concentrations were essential metals (Zn and Cu),
235 but relevant concentrations of potentially toxic elements as Hg and Se were observed too. Feather's
236 content and deposition rates of Co were negligible. Gradient for deposition in **feathers** was as
237 follows: $Zn > Cu > Hg > Se > Cr > Ni > Pb > As = Cd = V > Co$.

238 3.3 Gender differences in Kentish plover

239 Significant difference between males and females of KP were detected only for As (Mann-Whitney
240 U-test: $Z = -2.103$, $p = 0.041$), with females characterized by **As** feather concentration about 3 times
241 higher than males. No significant differences were observed for the other trace elements (Table 1).
242 Plots showing the variability of the power as a function of the difference in concentration between

243 sexes and standard deviation are reported in Supplementary Material, and allowed for an estimation
244 of the power achieved with available data.

245 3.4 Correlation among trace elements in Kentish plover

246 Several positive intercorrelations among trace elements occurred in KP feathers (Table 2). In
247 particular, Co, Se, Pb V and Zn were strongly and positively correlated. Similarly, Cd was
248 correlated with Co, Se and V, but not with Pb and Zn. A significant correlation was also observed
249 between Cr and Cu, whilst for As, Hg, and Ni no significant correlation were detected.

250 No correlations between bird weight and trace element concentrations were observed.

251 **4 Discussion**

252 Accumulation of trace elements in feathers is a topic that was treated in several studies, but the
253 possible contribution of environmental contamination to the conservation status of the threatened
254 species *Charadrius alexandrinus* was never assessed before. The use of surrogate species to
255 estimate exposure in a study area, although it is a feasible and often exploited approach, should be
256 avoided, whenever possible. In facts, as stated by Miller et al. (2019) "surrogate species of varying
257 body mass and life history traits may not consistently reflect levels of exposure across *taxa*, or even
258 among congeners or sex-age classes in the same areas". Biomonitoring of contaminants should
259 focus on specific *taxa* of concern (Miller et al., 2019), especially when threatened and/or declining
260 species (as KP) are target of the study.

261 Although sample size may appear low, it should be noted that 13 specimens represent the 48% of
262 the nesting population of KP in the study area in 2018, and 43% of the nesting specimens censused in
263 2017. Studies involving Charadriidae are generally characterized by a limited sample size, also
264 when identification of gender-related differences is of concern (Lucia et al., 2014, 2012; see
265 Supplementary Material), probably due to the less gregarious behaviour of waders as compared
266 with Laridae and Anatidae. For these reasons, we believe that sample size is adequate to assess the

267 possible exposure of the local population of KP to trace elements. However, we are aware that the
268 power of the statistical analysis achieved with 13 individuals may be low to effectively explore
269 possible gender-related differences for some trace element, due to both sample size and variability
270 among individuals of the same sex.

271 Trace elements in feathers may derive both from local exposure to contaminants through food
272 ingestion and mobilization from internal tissues at the time of feather formation (Burger and
273 Gochfeld, 1991, 1996); for this reason it is fundamental to know when a given species moults and,
274 if a migratory species is of concern, where it is when it moults (Burger and Gochfeld, 1996).

275 In KP, the molt of tail feathers occurs in the post-nuptial period (Ginn and Melville, 1983). In most
276 of European migrant populations the molt generally occurs partly in breeding areas and partly in
277 staging/stopover area, after a first post-breeding migration (Newton, 2009). However, in areas
278 where KP is resident or short-distance migrant, the molt may occur in breeding grounds. Monitoring
279 and ring recoveries collected over a period of 4 years on birds ringed in the Cavallino-Treporti
280 peninsula confirmed that KP individuals spent also their post-nuptial and overwintering period in
281 the breeding area. Thus, trace elements accumulation in feathers may be ascribed to local exposure
282 to contaminants, although the lack of GPS tracking studies cannot exclude the occurrence of
283 exceptions to this main trend.

284 4.1 Gender differences in Kentish Plover

285 The distinct plumage of males and females during breeding allows for the investigation of sex-
286 related differences in accumulation of trace elements. The present study represents the first
287 experience with KP on this topic.

288 Statistical analysis highlighted that significant differences between males and females occurred only
289 for As. Nevertheless, the power of the analysis may be too low to allow for an accurate assessment
290 of such differences for most of the analyzed trace elements. In any case, literature data concerning

291 **Charadriidae** suggested that gender-related differences in trace element accumulation in feathers are
292 an exception rather than a rule. In fact, Lucia et al. (2014, 2012) observed differences only for Cu
293 in dunlin (*Calidris alpina*) and Hg in redshank (*T. totanus*), whilst Ashbaugh et al. (2018) reported
294 significantly higher concentrations of Se in female snowy plovers (*C. nivosus*) than in males.

295 As for As, Lucia et al. (2010) observed that sex was a significant factor only for greylag goose
296 (*Anser anser*), with females characterized by higher burdens than males.

297 Little is known about As in birds, also regarding intraspecific factors that may influence
298 accumulation as gender, age and size (Sánchez-Virosta et al., 2015). Some authors pointed out that
299 As - and trace element concentrations in general - may differ when size dimorphism is relevant
300 (Janssens et al., 2001), but this hypothesis seems to be not applicable to KP, due to the minimal size
301 and weight differences between males and females. Ashbaugh et al. (2018) postulated that higher
302 concentrations of Se in females of *C. nivosus* may be due to sex-based susceptibility to uptake or
303 metabolic differences between sexes, but no evidence supporting these hypotheses were adduced.
304 Dissimilarity in prey type may contribute to differential accumulation in females and males
305 (Janssens et al., 2001), **since** type of food eaten plays a predominant role in the As amounts that
306 birds may accumulate (Lebedeva, 1997). Further studies on diet of KP, possibly involving stable
307 isotope analysis, are needed to determine whether differences in As concentrations may be due to
308 different diet composition between males and females, or whether other variables should be taken
309 into consideration, including gender-related metabolic differences.

310 4.2 Correlation among trace elements in KP tail feathers

311 Spearman's correlations evidenced that accumulation of essential elements in tail feathers of KP is
312 often positively correlated with deposition of **non-essential** metals. These correlations suggest that
313 **non-essential** metals (i.e. Pb and Cd) may be taken up through the same mechanism that allow the

314 assimilation of essential trace elements (including Se, Zn and Cu), and/or that essential and toxic
315 elements may come from same prey items.

316 Except for Hg (Dominik et al., 2014), there is a general lack of information concerning the transfer
317 of contaminants along the benthic food web of the Venice Lagoon, so it is not possible to identify
318 whether these correlations are effectively due to contaminant transfer from sediments to biota.

319 Comparison with literature data is of little help in this context, since too many variables differed
320 among studies, including contamination levels in study areas, feeding habits of the species, type of
321 feather, and trace elements analyzed. Consequently, correlation data are often not consistent and, in
322 some cases, also contrasting. As an example, Karimi et al. (2016) observed correlations between Pb
323 and Zn in primary and tails feathers of waterfowls in Iran (as in the present study), but same
324 elements were not correlated in waterfowls **from** wetlands of Korea (Kim and Oh, 2014). Similarly,
325 Pb and Cd were positively correlated in seabirds as pigeon guillemots *Cephus columba* and flesh-
326 footed shearwaters (*Puffinus carneipes*) (Bond and Lavers, 2011; Burger et al., 2007), but not in KP
327 (present work), Laughing gull *Larus atricilla*, and waterfowls (Gochfeld et al., 1996b; Karimi et al.,
328 2016). Correlation between Cu and Cr were not observed in previous studies.

329 Interestingly, data concerning KP highlighted also the lack of correlation between Se and Hg, a
330 condition often observed in birds (Ashbaugh et al., 2018; Burger et al., 2007). Since Se is reported
331 to mitigate the effects due to Hg in birds (Burger et al., 1993; Ohlendorf and Heinz, 2011), the lack
332 of this correlation may be symptomatic of an increased risk of **Hg-related** toxicity towards KP,
333 especially for those specimens **characterized by** higher Hg/Se molar concentration ratio.

334 4.3 Ecotoxicological significance of trace element levels for Kentish plover

335 **V**ery little is known concerning accumulation of trace elements in KP, **so the** assessment of the
336 significance of trace elements accumulated in KP feathers was based on: **1)** available adverse-
337 effects thresholds for birds, and **2)** comparison with literature data regarding waders and seabirds

338 sampled in different geographical areas. These latter data are reported in detail in *Supplementary*
339 *Material*.

340 Adverse-effects thresholds are available for main elements known to exert toxicity towards birds
341 (Hg, Cd and Pb) and Se. Besides these elements, there is a general lack of laboratory experiments
342 relating feather concentrations with adverse effects; other available thresholds (i.e. for Cr, Zn and
343 Ni) are most often values empirically derived from internal tissue concentrations.

344 **4.3.1 Mercury**

345 The trace element causing major concern for the conservation of KP is surely Hg. Feathers are
346 known to represent a major pathway for birds to eliminate the accumulated Hg, as they may contain
347 up to 70-90% of the total Hg body burden (Braune and Gaskin, 1987; Honda et al., 1986). Relevant
348 Hg concentrations were thus expected for KP, also in consideration of its invertivorous feeding
349 habits (Castro et al., 2009; Cramp and Simmons, 1983) and evidences that benthic macrofauna of
350 Venice lagoon accumulates Hg to a significant extent (Dominik et al., 2014). Nevertheless, it was
351 unexpected to measure Hg concentrations above the toxicity threshold of 5000 $\mu\text{g kg}^{-1}$ (Eisler,
352 2000, 1987) in 11 out of 13 analyzed KP. **This outcome is critical and of great ecological relevance:**
353 **even if sample size is relatively small, these data underline that a relevant part of the nesting**
354 **population (at least 39%) was surely exposed to Hg at levels that may have detrimental effects on**
355 **its conservation. In facts,** according to Eisler (2000), feather concentrations above 5000 ppb are
356 symptom of a severe exposure to Hg, and are associated with possible impairments in reproduction
357 and sterility in a number of species. In facts, dietary exposure to Hg or methyl-mercury does not
358 affect survival of adults **birds** (Whitney and Cristol, 2017), but may depress reproductive success by
359 reducing the clutch size (Tartu et al., 2013), reducing the number of hatched and fledged chicks per
360 brood (Hallinger and Cristol, 2011; Taylor and Cristol, 2015; Varian-Ramos et al., 2014),
361 increasing latency to renest after loss of a clutch (Varian-Ramos et al., 2014) or changing endocrine
362 expression, reproductive behavior and partner selection (Jayasena et al., 2011). All these sub-lethal

363 effects may significantly contribute to the depletion of the population at local scale, by reducing
364 hatching and survival of hatchlings and fledglings.

365 Moreover, Hg concentrations measured in KP are among the highest reported in the literature for
366 water birds. Feather concentrations similar or higher than those observed in Venice Lagoon were
367 reported only for some piscivorous species as cormorants, terns, and petrels (Bond and Lavers,
368 2011; Eagles-Smith et al., 2008; Hribšek et al., 2017; Monteiro et al., 1995; Ochoa-acunã et al.,
369 2002). In waders, Hg occur at levels generally lower than those observed for KP, also in
370 contaminated areas as the Derwent estuary, Australia (Einoder et al., 2018). **An exception to this**
371 **trend is the** Black-necked stilt *Himantopus mexicanus* (Eagles-Smith et al., 2008).

372 These data strengthen the hypothesis that exposure to inorganic Hg and/or monomethyl-Hg in the
373 Venice Lagoon is of concern, not only for KP but also for other **species** feeding on macrobenthic
374 community. Although the available data do not allow to establish whether reproduction failures
375 observed in the Cavallino-Treporti peninsula, such as increased number of unfertilized and
376 unhatched eggs (Picone, unpublished data) may be due to dietary exposure to Hg, they indicate that
377 trophic transfer of Hg needs a continuous monitoring.

378 **4.3.2 Cadmium**

379 In birds, Cd taken up through food ingestion is mainly deposited in kidney and liver, and only
380 secondarily in feathers. According to Burger and Gochfeld (2000), toxic effects of Cd towards
381 seabirds may occur at feather concentration ranging from 100 up to 2,000 $\mu\text{g kg}^{-1}$. These thresholds,
382 however, were empirically derived from adverse-effect concentrations in kidney obtained in
383 laboratory (10,000 $\mu\text{g kg}^{-1}$ of Cd) (Eisler, 2000) and kidney:feathers ratios in pelagic birds. **Thus,**
384 **these thresholds** should be considered as guidance values rather than adverse-effect concentrations.
385 In any case, average concentration in KP exceeds 100 $\mu\text{g kg}^{-1}$, suggesting that **some individuals**
386 **may have been** exposed to Cd to an extent that may lead to sub-lethal effects including growth

387 inhibition, disruption of Ca metabolism, anemia, renal and testicular damages (Burger, 2008;
388 Burger and Gochfeld, 2000; Eisler, 2000 and citations herein).

389 Concentrations of Cd **measured** in KP **in the present work** are lower than those reported by Kim and
390 Koo (2008) for the same species ($910 \mu\text{g kg}^{-1}$), and also than those measured in ducks and waders in
391 Iran (Karimi et al., 2016) and Korea (Kim and Oh, 2012). Nevertheless, Cd levels in KP feathers are
392 considerably higher than those observed in waders and ducks collected in wetlands reported to be
393 contaminated by Cd and/or other elements in France and Australia (Einoder et al., 2018; Lucia et
394 al., 2010).

395 Based on a liver-to-kidney ratio, Lucia et al. (2010) have estimated a low-chronic exposure to Cd
396 for graylag goose (*Anser anser*), red knot (*Calidris canutus*) and grey plover (*Pluvialis squatarola*),
397 with minor threat for survival. Since mean Cd concentrations in **feathers of** these birds ranged in the
398 interval of $40\text{-}90 \mu\text{g kg}^{-1}$ (Lucia et al. 2010), according to these data the exposure of KP to Cd may
399 be considered from low to moderate, suggesting that long term effects cannot be excluded, as also
400 evidenced by the comparison with the adverse-effect threshold.

401 **4.3.3 Lead**

402 **KP exposure to Pb in the Venice Lagoon seems to be negligible and occurrence of toxicosis**
403 **unlikely.** Concentrations of Pb in tail feathers of KP are by far lower than the $4,000 \mu\text{g kg}^{-1}$ toxicity
404 threshold reported by Burger and Gochfeld (2000), associated with drooped wings, loss of appetite,
405 lethargy, weakness, impaired locomotion, lowered reproductive success and survival (Burger and
406 Gochfeld, 2004). Kim and Koo (2008) for the same species observed an average Pb concentration
407 of $9,840 \mu\text{g kg}^{-1}$, one order of magnitude higher than that measured in the present work, and more
408 than twice the adverse-effect threshold reported by Burger and Gochfeld (2000).

409 In general, Pb concentrations **measured in this work** are among the lowest reported in bird's
410 feathers. Transfer of Pb through the trophic web seems not to be a factor for conservation of KP.

411 4.3.4 Selenium

412 Selenium is essential in ultra-trace, but may become toxic at concentrations slightly higher than
413 those required for homeostasis (Ohlendorf and Heinz, 2011). Background concentrations in feathers
414 are reported to be typically less than 2,000 $\mu\text{g kg}^{-1}$ but a range of 1,000-4,000 $\mu\text{g kg}^{-1}$ is most often
415 considered, **due to interspecific variability** (Ohlendorf and Heinz, 2011). A Se concentration of
416 5,000 $\mu\text{g kg}^{-1}$ was identified as a provisional threshold of toxicity by a number of researchers
417 (Ashbaugh et al., 2018; Burger et al., 1993; Ohlendorf and Heinz, 2011), although concentrations
418 that may be directly related to impairments in adults or offspring are not available. Two specimens
419 of KP exceeded the 5,000 $\mu\text{g kg}^{-1}$ threshold, suggesting that possible adverse effects due to Se
420 cannot be excluded. Nevertheless, in migratory waders sampled at middle latitudes (including the
421 KP's sister species *C. nivosus*), Se occurs often at concentrations higher than 5,000 $\mu\text{g kg}^{-1}$
422 (Ashbaugh et al., 2018; Burger et al., 1993; Lucia et al., 2010; St. Clair et al., 2015). **Thus**, it cannot
423 be excluded that background concentration in Charadriidae and other shorebirds may be naturally
424 higher than those observed in seabirds and arctic birds. Further research is needed to establish
425 background values for waders and identify adverse-effect threshold for Se.

426 4.3.5 Chromium

427 Burger and Gochfeld (2000) suggested that impairments due to Cr may occur at feather
428 concentrations above 2,800 $\mu\text{g kg}^{-1}$. Since mean concentration in tail feathers of KP is lower than
429 this value, and only a single female (2,745 $\mu\text{g kg}^{-1}$) approximates this adverse-effect concentration
430 derived from liver concentrations, Cr does not seem **to be of concern** for the conservation status of
431 KP in the Venice Lagoon.

432 As compared with other Charadriidae, feather concentrations of Cr in KP are low; more difficult is
433 the comparison with other families, since concentrations are very variable among sites and species.

434 **4.3.6 Nickel**

435 Nickel is an essential metal required by organisms for growth; nevertheless, bird exposed to high
436 concentrations of this element through diet may suffer adverse effects including metabolic upset,
437 altered bone density, inhibition of growth (Eisler, 2000). Literature data showed that KP feather
438 concentrations are slightly higher than levels occurring in a number of seabird species. However,
439 exposure to Ni in the Venice Lagoon appears to be not of concern since the adverse-effect threshold
440 proposed by Outridge and Scheuhammer (1993) is $4000 \mu\text{g kg}^{-1}$, about 3 times higher than the
441 average concentration measured for KP.

442 **4.3.7 Zinc**

443 Zinc is an essential element for birds, playing a key role in enzymatic reactions, bone structure and
444 growth and for cellular immunity (Park et al., 2004). Nevertheless, it may become toxic at
445 concentrations overwhelming homeostasis. In the case of Zn, the occurrence of pigments able to
446 bind metals as eumelanin (Klasing, 1998; Lodenius and Solonen, 2013) may increase the feather
447 concentration, providing "false positive" responses concerning exposure to and accumulation of Zn.

448 This is not the case of KP. **Firstly, the sixth rectrix of KP lacks of pigmentation.** Secondly, the
449 adverse-effect thresholds for Zn reported by Einoder et al. (2018) - based on data provided by
450 Gasaway and Buss (1972) - is at $200,000 \mu\text{g kg}^{-1}$, a level more than 3 times higher than the average
451 Zn concentration measured in KP tail feathers. **Thirdly,** Zn concentrations in feathers of KP are
452 among the lowest observed in literature. Hazard due to Zn is thus improbable in Venice Lagoon for
453 KP.

454 **4.3.8 Other trace elements (As, Co, Cu and V)**

455 Although As is carcinogen and may cause behavioral and physical effects in birds, fate and
456 toxicology of As in water birds are poorly known, and consequently no theoretical nor empirical
457 adverse-effect threshold is available (Bond and Lavers, 2011; Burger et al., 2007). In any case,
458 feathers are not expected to be a major structure for As deposition in birds (Fujihara et al., 2004).

459 Concentrations of As in KP are the lowest observed for Charadriidae, with a mean value about one
460 order of magnitude lower than the concentrations measured in dunlin (*Calidris alpina*) and grey
461 plover (*P. squatarola*) (Goede et al., 1989; Lucia et al., 2010). On the basis of these data, hazard
462 due to exposure to As in the area of Venice Lagoon seems to be negligible for KP.

463 Information concerning effects of Co towards birds are scarce, also due to the lack of controlled
464 laboratory studies relating Co concentrations in feathers and other tissue with adverse effects
465 (Burger et al., 2018). However, average concentrations observed in KP are among the lowest
466 observed in water birds for Co; most probably this element **does not pose** any risk to conservation
467 of KP in Venice Lagoon area.

468 Due to its essential role in many biochemical and enzymatic reactions, Cu tends to be rather
469 uniform among bird species, and deviations from this trend are most often due to some species-
470 specific bioaccumulation pattern rather than to habitat contamination (Kim et al., 1996). According
471 to Eisler (2000), in birds from contaminated sites (near smelters) Cu in feathers usually ranges in
472 the interval 23,000 - 53,000 $\mu\text{g kg}^{-1}$. **Feather concentrations in KP are clearly lower than this values**
473 **(only a single male approximates 20,000 $\mu\text{g kg}^{-1}$), suggesting a low exposure to Cu. Nevertheless,**
474 **predictions concerning possible toxic effects are not feasible with the available data. Feather**
475 **concentrations of Cu measured in KP are** consistent with **values** reported for other Charadriidae and
476 waterbirds, but higher than **those** observed by Kim and Koo (2008) **for the same species** in Korea.

477 Deposition of V in feathers has been largely disregarded by ecotoxicologists, although there are
478 evidences of vanadium-induced toxicity in birds, including gastrointestinal hemorrhage, loss of
479 weight and hearth failures (Rattner et al., 2006). However, in the case of KP, **V** is not expected to be
480 of concern: feather concentrations are low as compared with shearwaters and waterfowls (Bond and
481 Lavers, 2011; Hosseini Alhashemi et al., 2011). **Moreover**, uptake of sediment-bound V from
482 rooted plants seems to be a preferential pathway for V transfer into the food web (Hosseini
483 Alhashemi et al., 2011).

484 **4.3.9 Daily deposition rates (DDRs)**

485 The present work is the first experience with KP and also one of the few studies concerning daily
486 deposition rates in birds. Normalizing the content of contaminant by age of feather is potentially
487 more biologically meaningful than reporting data normalized to unit of feather weight, as stated by
488 Clarkson and Riscassi (2011). These authors agree with the theory of Bortolotti (2010) and argue
489 that potential biological impacts of ingested contaminants may be masked in a healthy individual
490 with structurally dense feathers, as the normalization per unit mass of feather will be low, whilst
491 analyzing a feather's content by daily deposition may potentially correct for this effect. Moreover,
492 determining the average daily deposition into growing feathers provide a higher sensitivity for
493 detecting differences among species than normalizing the content of trace elements per unit of
494 feather weight (Clarkson and Riscassi, 2011). Nevertheless, although its relevance, this approach
495 was rarely applied in ecotoxicology.

496 Deposition rate of Hg in KP resulted lower than the DDR observed in nestlings of double-crested
497 cormorant *Phalacrocorax auritus* (12 ng d⁻¹) but higher than that measured in the glossy ibis
498 *Plegadis falcinellus* (2 ng d⁻¹) along the coasts of New York and Virginia (Clarkson and Riscassi,
499 2011). A lower DDR as compared with double-crested cormorant was expected, due to the strictly
500 piscivorous feeding habits of *P. auritus*. On the other hand, a deposition rate about 2 times higher
501 than chicks of glossy ibis, that feed mostly on aquatic invertebrates and insects as KP, is of concern
502 and evidences a significant exposure to Hg.

503 For elements other than Hg, deposition rate are not available in literature, so further comparisons
504 cannot be performed. However, some inferences can be made. High deposition rates observed for
505 essential elements as Cu and Zn are most probably due to their structural role in the keratinous
506 composition of the feather (Howell et al., 2017). Due to the lack of pigmentation in the sixth rectrix
507 of KP, deposition rates of Zn and Cu cannot be related with the occurrence of eumelanin or
508 pheomelanin, known to serve as metal reservoir also in feathers (Edwards et al., 2016). Low

509 deposition rate observed for Co indicates negligible blood concentration during feather growth and
510 thus a limited availability of Co through the food web. Deposition rates for As, Cd and V are quite
511 low, as compared with other trace elements. However, in the case of Cd, despite a low DDR the
512 feather concentration exceeded the adverse-effect threshold of $100 \mu\text{g kg}^{-1}$, indicating that DDR
513 alone cannot be used as a predictor of metal-related toxic effects.

514 Further studies are needed to acquire more data concerning DDR in a number of species with
515 different feeding habits, in order to obtain benchmarks to estimate exposure.

516 **5 Conclusions**

517 This work represents the first study aimed at identifying possible disturbances related to
518 environmental contamination on the conservation status of the threatened species *Charadrius*
519 *alexandrinus*, by using a non-invasive approach. Although the sample size was relatively small, it
520 allowed to gain an insight into possible harmful exposure of the local population of KP to trace
521 elements, especially as concern the identification of main threats for the species.

522 Mercury emerged as a major threat to the conservation of the species at local scale; feather
523 concentrations above the adverse-effect threshold in 11 out of 13 analyzed birds underlined a
524 probable risk of Hg-related toxicosis, that may lead to impairments in the reproductive success of
525 KP, and of other water birds with similar feeding habits. Further studies are needed to verify
526 whether reproductive failures already observed in the field may be linked to Hg-related toxicosis or
527 other causes.

528 Even if to a lesser extent than Hg, also exposure to Cd and Se may be of concern for KP
529 conservation. For these elements there is the need to identify which are the compartments of the
530 benthic food web along which the toxicant may be transferred from the sediment to the birds (i.e.
531 polychaetes, amphipods, small bivalves, etc.).

532 Daily deposition rates provided a measure of the exposure to contaminants that may be more
533 biologically meaningful than the mass/mass ratio and can provide useful insights concerning the
534 transfer of contaminants along the food web that may be directly related with dietary intakes in an
535 ecological risk assessment approach.

536 **6 Funding sources**

537 This work was partly supported by the Cavallino-Treporti municipality (det. 355/2017 and det.
538 1552/2018).

539

540 **7 References**

- 541 Adout, A., Hawlena, D., Maman, R., Paz-Tal, O., Karpas, Z., 2007. Determination of trace elements
542 in pigeon and raven feathers by ICPMS. *Int. J. Mass Spectrom.* 267, 109–116.
543 doi:<https://doi.org/10.1016/j.ijms.2007.02.022>
- 544 Argüelles-Ticó, A., Küpper, C., Kelsh, R.N., Kosztolányi, A., Székely, T., van Dijk, R.E., 2016.
545 Geographic variation in breeding system and environment predicts melanin-based plumage
546 ornamentation of male and female Kentish plovers. *Behav. Ecol. Sociobiol.* 70, 49–60.
547 doi:[10.1007/s00265-015-2024-8](https://doi.org/10.1007/s00265-015-2024-8)
- 548 Ashbaugh, H.M., Conway, W.C., Haukos, D.A., Collins, D.P., Comer, C.E., French, A.D., 2018.
549 Evidence for exposure to selenium by breeding interior snowy plovers (*Charadrius nivosus*) in
550 saline systems of the Southern Great Plains. *Ecotoxicology* 27, 703–718. doi:[10.1007/s10646-](https://doi.org/10.1007/s10646-018-1952-2)
551 [018-1952-2](https://doi.org/10.1007/s10646-018-1952-2)
- 552 Blus, L.J., Henny, C.J., Hoffman, D.J., Grove, R.A., 1995. Accumulation in and effects of lead and
553 cadmium on waterfowl and passerines in northern Idaho. *Environ. Pollut.* 89, 311–318.
554 doi:[10.1016/0269-7491\(94\)00069-P](https://doi.org/10.1016/0269-7491(94)00069-P)
- 555 Bond, A.L., Lavers, J.L., 2011. Trace element concentrations in feathers of Flesh-footed
556 Shearwaters (*Puffinus carneipes*) from across their breeding range. *Arch. Environ. Contam.*
557 *Toxicol.* 61, 318–326. doi:[10.1007/s00244-010-9605-3](https://doi.org/10.1007/s00244-010-9605-3)
- 558 Bortolotti, G.R., 2010. Flaws and pitfalls in the chemical analysis of feathers: bad news — good
559 news for avian chemoecology and toxicology. *Ecol. Appl.* 20, 1766–1774. doi:[10.1890/09-](https://doi.org/10.1890/09-1473.1)
560 [1473.1](https://doi.org/10.1890/09-1473.1)
- 561 Braune, B.M., Gaskin, D.E., 1987. A mercury budget for the Bonaparte's gull during autumn moult.
562 *Ornis Scand.* 18, 244. doi:[10.2307/3676891](https://doi.org/10.2307/3676891)

563 Burger, J., 2008. Assessment and management of risk to wildlife from cadmium. *Sci. Total*
564 *Environ.* 389, 37–45. doi:10.1016/J.SCITOTENV.2007.08.037

565 Burger, J., 1993. Metals in avian feathers: bioindicators of environmental pollution. *Rev. Environ.*
566 *Toxicol.* 5, 203–311.

567 Burger, J., Gochfeld, M., 2004. Marine Birds as Sentinels of Environmental Pollution. *Ecohealth* 1,
568 263–274. doi:10.1007/s10393-004-0096-4

569 Burger, J., Gochfeld, M., 2000a. Metals in Albatross Feathers from Midway Atoll: Influence of
570 Species, Age, and Nest Location. *Environ. Res.* 82, 207–221. doi:10.1006/enrs.1999.4015

571 Burger, J., Gochfeld, M., 2000b. Metal levels in feathers of 12 species of seabirds from Midway
572 Atoll in the northern Pacific Ocean. *Sci. Total Environ.* 257, 37–52. doi:10.1016/S0048-
573 9697(00)00496-4

574 Burger, J., Gochfeld, M., 1997. Risk, mercury levels, and birds: relating adverse laboratory effects
575 to field biomonitoring. *Environ. Res.* 75, 160–172. doi:10.1006/ENRS.1997.3778

576 Burger, J., Gochfeld, M., 1996. Heavy metal and selenium levels in Franklin’s Gull (*Larus*
577 *pipixcan*) parents and their eggs. *Arch. Environ. Contam. Toxicol.* 30, 487–491.
578 doi:10.1007/s002449900067

579 Burger, J., Gochfeld, M., 1991. Lead, mercury, and cadmium in feathers of tropical terns in Puerto
580 Rico and Australia. *Arch. Environ. Contam. Toxicol.* 21, 311–315. doi:10.1007/BF01055351

581 Burger, J., Gochfeld, M., Sullivan, K., Irons, D., 2007. Mercury, arsenic, cadmium, chromium lead,
582 and selenium in feathers of pigeon guillemots (*Cepphus columba*) from Prince William Sound
583 and the Aleutian Islands of Alaska. *Sci. Total Environ.* 387, 175–184.
584 doi:10.1016/J.SCITOTENV.2007.07.049

- 585 Burger, J., Mizrahi, D., Tsipoura, N., Jeitner, C., Gochfeld, M., 2018. Mercury, lead, cadmium,
586 cobalt, arsenic and selenium in the blood of Semipalmated sandpipers (*Calidris pusilla*) from
587 Suriname, South America: Age-related differences in wintering site and comparisons with a
588 stopover site in New Jersey, USA. *Toxics* 6. doi:10.3390/toxics6020027
- 589 Burger, J., Seyboldt, S., Morganstein, N., Clark, K., 1993. Heavy metals and selenium in feathers of
590 three shorebird species from Delaware bay. *Environ. Monit. Assess.* 28, 189–198.
591 doi:10.1007/BF00547037
- 592 Castro, M., Masero, J.A., Pérez-Hurtado, A., Amat, J.A., Megina, C., 2009. Sex-related seasonal
593 differences in the foraging strategy of the Kentish plover. *Condor* 111, 624–632.
594 doi:10.1525/cond.2009.080062
- 595 Cecconi, G., Nascimbeni, P., 1997. Ricostruzione e naturalizzazione delle dune artificiali sul
596 litorale del Cavallino.
- 597 Clarkson, C.E., Riscassi, A., 2011. Using ptilochronology to determine daily mercury deposition in
598 feathers of nestling waterbirds. *Environ. Toxicol. Chem.* 30, 2081–2083. doi:10.1002/etc.591
- 599 Cramp, S., Simmons, K.E.L., 1983. Handbook of the birds of Europe, the Middle East and North
600 Africa. The birds of the Western Palearctic: 3. Waders to gulls. Oxford University Press,
601 Oxford.
- 602 Dauwe, T., Bervoets, L., Blust, R., Pinxten, R., Eens, M., 2000. Can excrement and feathers of
603 nestling songbirds be used as biomonitors for heavy metal pollution? *Arch. Environ. Contam.*
604 *Toxicol.* 39, 541–546. doi:10.1007/s002440010138
- 605 Delany, S., Scott, D.A., Dodman, T., Stroud, D.A., 2009. The wader atlas: An atlas of wader
606 populations in Africa and western Eurasia. Wetlands International, Wageningen.
- 607 Del Vecchio, S., Pizzo, L., Buffa, G., 2015. The response of plant community diversity to alien

608 invasion: evidence from a sand dune time series. *Biodivers. Conserv.* 24, 371–392.
609 doi:10.1007/s10531-014-0814-3

610 Domínguez, J., Vidal, M., 2003. Influencia del investigador en el éxito reproductivo del Chorlito
611 Patinegro *Charadrius alexandrinus*. *Ardeola* 50, 15–19.

612 Dominik, J., Tagliapietra, D., Bravo, A.G., Sigovini, M., Spangenberg, J.E., Amouroux, D., Zonta,
613 R., 2014. Mercury in the food chain of the Lagoon of Venice, Italy. *Mar. Pollut. Bull.* 88, 194–
614 206. doi:10.1016/J.MARPOLBUL.2014.09.005

615 Donazzolo, R., Merlin, O.H., Vitturi, L.M., Orio, A.A., Pavoni, B., Perin, G., Rabitti, S., 1981.
616 Heavy metal contamination in surface sediments from the Gulf of Venice, Italy. *Mar. Pollut.*
617 *Bull.* 12, 417–425. doi:10.1016/0025-326X(81)90160-0

618 Eagles-Smith, C.A., Ackerman, J.T., Adelsbach, T.L., Takekawa, J.Y., Miles, A.K., Keister, R.A.,
619 2008. Mercury correlations among six tissues for four waterbird species breeding in San
620 Francisco Bay, California, USA. *Environ. Toxicol. Chem.* 27, 2136. doi:10.1897/08-038.1

621 Edwards, N.P., van Veelen, A., Anné, J., Manning, P.L., Bergmann, U., Sellers, W.I., Egerton,
622 V.M., Sokaras, D., Alonso-Mori, R., Wakamatsu, K., Ito, S., Wogelius, R.A., 2016. Elemental
623 characterisation of melanin in feathers via synchrotron X-ray imaging and absorption
624 spectroscopy. *Sci. Rep.* 6, 34002. doi:10.1038/srep34002

625 Einoder, L.D., MacLeod, C.K., Coughanowr, C., 2018. Metal and Isotope Analysis of Bird Feathers
626 in a Contaminated Estuary Reveals Bioaccumulation, Biomagnification, and Potential Toxic
627 Effects. *Arch. Environ. Contam. Toxicol.* 75, 96–110. doi:10.1007/s00244-018-0532-z

628 Eisler, R., 2000. *Handbook of Chemical Risk Assessment*. CRC Press. doi:10.1201/9781420032741

629 Eisler, R., 1987. Mercury hazards to fish, wildlife, and invertebrates: A synoptic review, *Fish and*
630 *Wildlife Service Biological Report* 85 (1.10).

- 631 Fraga, R.M., Amat, J.A., 1996. Breeding biology of a Kentish plover (*Charadrius alexandrinus*)
632 population in an inland saline lake. *Ardeola* 43, 69–85.
- 633 Fujihara, J., Kunito, T., Kubota, R., Tanaka, H., Tanabe, S., 2004. Arsenic accumulation and
634 distribution in tissues of black-footed albatrosses. *Mar. Pollut. Bull.* 48, 1153–1160.
635 doi:10.1016/J.MARPOLBUL.2004.03.007
- 636 Furness, R.W., 1996. Cadmium in Birds, in: Beyer, W.N., Heinz, G.H., Redmon-Norwood, A.W.
637 (Eds.), *Environmental Contaminants in Wildlife: Interpreting Tissue Concentrations*. Lewis
638 Publishers, Boca Raton, FL, pp. 389–404.
- 639 García-Fernández, A.J., Espín, S., Martínez-López, E., 2013. Feathers as a biomonitoring tool of
640 polyhalogenated compounds: A review. *Environ. Sci. Technol.* 47, 3028–3043.
641 doi:10.1021/es302758x
- 642 Gasaway, W.C., Buss, I.O., 1972. Zinc Toxicity in the Mallard Duck. *J. Wildl. Manage.* 36, 1107.
643 doi:10.2307/3799239
- 644 Ginn, H.B., Melville, D.S., 1983. *Moult in birds*. British Trust for Ornithology.
- 645 Gochfeld, M., Belant, J.L., Shukla, T., Benson, T., Burger, J., 1996. Heavy metals in laughing gulls:
646 gender, age and tissue differences, *Environmental Toxicology and Chemistry*.
- 647 Goede, A., Nygard, T., De Bruin, M., Steinnes, E., 1989. Selenium, mercury, arsenic and cadmium
648 in the life-cycle of the Dunlin, *Calidris alpina*, a migrant wader. *Sci. Total Environ.* Elsevier
649 *Sci. Publ. B.V* 78, 205–218.
- 650 Grubb, T.C.J., 2006. *Ptilochronology. Feather time and the biology of birds*. Oxford University
651 Press.
- 652 Grubb, T.C.J., 1989. Ptilochronology: feather growth bars as indicators of nutritional status. *Auk*

653 106, 314–320.

654 Hallinger, K.K., Cristol, D.A., 2011. The role of weather in mediating the effect of mercury
655 exposure on reproductive success in tree swallows. *Ecotoxicology* 20, 1368–1377.
656 doi:10.1007/s10646-011-0694-1

657 Honda, K., Nasu, T., Tatsukawa, R., 1986. Seasonal changes in mercury accumulation in the black-
658 eared kite, *Milvus migrans lineatus*. *Environ. Pollut. Ser. A, Ecol. Biol.* 42, 325–334.
659 doi:10.1016/0143-1471(86)90016-4

660 Hosseini Alhashemi, A.S., Karbassi, A.R., Hassanzadeh Kiabi, B., Monavari, S.M., Nabavi,
661 S.M.B., Sekhavatjou, M.S., 2011. Bioaccumulation of trace elements in trophic levels of
662 wetland plants and waterfowl birds. *Biol. Trace Elem. Res.* 142, 500–516.
663 doi:10.1007/s12011-010-8795-x

664 Hribšek, I., Jovičić, K., Karadžić, B., Skorić, S., 2017. Allocation of metals and trace elements in
665 different tissues of piscivorous species *Phalacrocorax carbo*. *Arch. Environ. Contam. Toxicol.*
666 73, 533–541. doi:10.1007/s00244-017-0452-3

667 Janssens, E., Dauwe, T., Bervoets, L., Eens, M., 2001. Heavy metals and selenium in feathers of
668 great tits (*Parus major*) along a pollution gradient. *Environ. Toxicol. Chem.* 20, 2815–2820.
669 doi:10.1002/etc.5620201221

670 Jaspers, V.L.B., Voorspoels, S., Covaci, A., Eens, M., 2006. Can predatory bird feathers be used as
671 a non-destructive biomonitoring tool of organic pollutants? *Biol. Lett.* 2, 283–285.
672 doi:10.1098/rsbl.2006.0450

673 Jayasena, N., Frederick, P.C., Larkin, I.L. V, 2011. Endocrine disruption in white ibises (*Eudocimus*
674 *albus*) caused by exposure to environmentally relevant levels of methylmercury. *Aquat.*
675 *Toxicol.* 105, 321–327. doi:https://doi.org/10.1016/j.aquatox.2011.07.003

- 676 Karimi, M.H.S., Hassanpour, M., Pourkhabbaz, A.R., Błaszczuk, M., Paluch, J., Binkowski, Ł.J.,
677 2016. Trace element concentrations in feathers of five Anseriformes in the south of the
678 Caspian Sea, Iran. *Environ. Monit. Assess.* 188, 1–7. doi:10.1007/s10661-015-5015-3
- 679 Kim, E.-Y., Ichihashi, H., Saeki, K., Atrashkevich, G., Tanabe, S., Tatsukawa, R., 1996. Metal
680 accumulation in tissues of seabirds from Chaun, northeast Siberia, Russia. *Environ. Pollut.* 92,
681 247–252. doi:10.1016/0269-7491(96)00007-3
- 682 Kim, J., Koo, T.-H., 2008. Heavy Metal Concentrations in Feathers of Korean Shorebirds. *Arch.*
683 *Environ. Contam. Toxicol.* 55, 122–128. doi:10.1007/s00244-007-9089-y
- 684 Kim, J., Oh, J.-M., 2014a. Concentration of trace elements in feathers of waterfowl, Korea.
685 *Environ. Monit. Assess.* 186, 8517–8525. doi:10.1007/s10661-014-4021-1
- 686 Kim, J., Oh, J.-M., 2014b. Concentration of trace elements in feathers of waterfowl, Korea.
687 *Environ. Monit. Assess.* 186, 8517–8525. doi:10.1007/s10661-014-4021-1
- 688 Kim, J., Oh, J.-M., 2012. Monitoring of heavy metal contaminants using feathers of shorebirds,
689 Korea. *J. Environ. Monit.* 14, 651. doi:10.1039/c2em10729e
- 690 Klasing, K.C., 1998. *Comparative avian nutrition*. Cab International.
- 691 Kosztolányi, A., Javed, S., Küpper, C., Cuthill, I.C., Shamsi, A. Al, Székely, T., 2009. Breeding
692 ecology of Kentish plover *Charadrius alexandrinus* in an extremely hot environment. *Bird*
693 *Study* 56, 244–252. doi:10.1080/00063650902792106
- 694 Lebedeva, N.V., 1997. Accumulation of heavy metals by birds in Southwest of Russia. *Russ. J.*
695 *Ecol.* 28, 41–46. doi:10.06
- 696 Lessells, C.M., 1984. The mating system of Kentish plovers *Charadrius alexandrinus*. *Ibis* (Lond.
697 1859). 126, 474–483.

- 698 Lodenius, M., Solonen, T., 2013. The use of feathers of birds of prey as indicators of metal
699 pollution. *Ecotoxicology* 22, 1319–1334. doi:10.1007/s10646-013-1128-z
- 700 Long, E.R., Ingersoll, C.G., MacDonald, D.D., 2006. Calculation and uses of mean sediment
701 quality guideline quotients: A critical review. *Environ. Sci. Technol.* 40, 1726–1736.
702 doi:10.1021/es058012d
- 703 Lucia, M., André, J.M., Gontier, K., Diot, N., Veiga, J., Davail, S., 2010. Trace element
704 concentrations (mercury, cadmium, copper, zinc, lead, aluminium, nickel, arsenic, and
705 selenium) in some aquatic birds of the southwest atlantic coast of France. *Arch. Environ.
706 Contam. Toxicol.* 58, 844–853. doi:10.1007/s00244-009-9393-9
- 707 Lucia, M., Bocher, P., Chambosse, M., Delaporte, P., Bustamante, P., 2014. Trace element
708 accumulation in relation to trophic niches of shorebirds using intertidal mudflats. *J. Sea Res.*
709 92, 134–143. doi:10.1016/J.SEARES.2013.08.008
- 710 Lucia, M., Bocher, P., Cosson, R.P., Churlaud, C., Bustamante, P., 2012. Evidence of species-
711 specific detoxification processes for trace elements in shorebirds. *Ecotoxicology* 21, 2349–
712 2362. doi:10.1007/s10646-012-0991-3
- 713 Markowski, M., Kaliński, A., Skwarska, J., Wawrzyniak, J., Bańbura, M., Markowski, J., Zieliński,
714 P., Bańbura, J., 2013. Avian feathers as bioindicators of the exposure to heavy metal
715 contamination of food. *Bull. Environ. Contam. Toxicol.* 91, 302–305. doi:10.1007/s00128-
716 013-1065-9
- 717 Miller, M.W.C., Lovvorn, J.R., Matz, A.C., Taylor, R.J., Latty, C.J., Brooks, M.L., Hollmén, T.E.,
718 2019. Interspecific patterns of trace elements in sea ducks: Can surrogate species be used in
719 contaminants monitoring? *Ecol. Indic.* 98, 830–839. doi:10.1016/J.ECOLIND.2018.11.023
- 720 Monteiro, L.R., Furness, R.W., Del Nevo, A.J., 1995. Environmental Contamination a n d

721 Toxicology Mercury Levels in Seabirds from the Azores, Mid-North Atlantic Ocean. Arch.
722 Environ. Contam. Toxicol 28, 304–309.

723 Newton, I., 2009. Moulting and plumage. Ringing Migr. 24, 220–226.
724 doi:10.1080/03078698.2009.9674395

725 Ochoa-acunã, H., Sepúlveda, M.S., Gross, T.S., 2002. Mercury in feathers from Chilean birds:
726 influence of location, feeding strategy, and taxonomic affiliation. Mar. Pollut. Bull. 44, 340–
727 345. doi:10.1016/S0025-326X(01)00280-6

728 Ohlendorf, H.M., Heinz, G.H., 2011. Selenium in Birds, in: Beyer, W.N., Meador, J.P. (Eds.),
729 Environmental Contaminants in Biota: Interpreting Tissue Concentrations, 2nd Edition. Lewis
730 Publishers, Boca Raton, FL, pp. 669–701.

731 Ohlendorf, H.M., Heinz, G.H., n.d. Selenium in Birds.

732 Outridge, P.M., Scheuhammer, A.M., 1993. Bioaccumulation and toxicology of nickel: implications
733 for wild mammals and birds. Environ. Rev. 1, 172–197. doi:10.1139/a93-013

734 Park, S.Y., Birkhold, S.G., Kubena, L.F., Nisbet, D.J., Ricke, S.C., 2004. Review on the role of
735 dietary zinc in poultry nutrition, immunity, and reproduction. Biol. Trace Elem. Res. 101, 147–
736 164. doi:10.1385/BTER:101:2:147

737 Perez-Hurtado, A., Goss-Custard, J.D., Garcia, F., 1997. The diet of wintering waders in Cádiz Bay,
738 southwest Spain. Bird Study 44, 45–52. doi:10.1080/00063659709461037

739 Picone, M., Bergamin, M., Delaney, E., Ghirardini, A.V., Kusk, K.O., 2018. Testing lagoonal
740 sediments with early life stages of the copepod *Acartia tonsa* (Dana): An approach to assess
741 sediment toxicity in the Venice Lagoon. Ecotoxicol. Environ. Saf. 147.
742 doi:10.1016/j.ecoenv.2017.08.042

- 743 Picone, M., Bergamin, M., Losso, C., Delaney, E., Arizzi Novelli, A., Ghirardini, A.V., 2016.
744 Assessment of sediment toxicity in the Lagoon of Venice (Italy) using a multi-species set of
745 bioassays. *Ecotoxicol. Environ. Saf.* 123. doi:10.1016/j.ecoenv.2015.09.002
- 746 R. Howell, N., Lavers, J.L., Uematsu, S., Paterson, D., Howard, D.L., Spiers, K., Jonge, M.D. de,
747 Hanley, T., Garrett, R., Banati, R.B., 2017. The Topobiology of Chemical Elements in Seabird
748 Feathers. *Sci. Rep.* 7, 1998. doi:10.1038/s41598-017-01878-y
- 749 Rattner, B.A., McKernan, M.A., Eisenreich, K.M., Link, W.A., Olsen, G.H., Hoffman, D.J.,
750 Knowles, K.A., McGowan, P.C., 2006. Toxicity and hazard of vanadium to Mallard ducks (
751 *Anas platyrhynchos*) and Canada geese (*Branta canadensis*). *J. Toxicol. Environ. Heal. Part*
752 *A* 69, 331–351. doi:10.1080/15287390500398265
- 753 Rondinini, C., Battistoni, A., Peronace, V., Teofili, C., 2013. Lista Rossa IUCN dei Vertebrati
754 Italiani. Roma.
- 755 Rothschild, R.F.N., Duffy, L.K., 2005. Mercury concentrations in muscle, brain and bone of
756 Western Alaskan waterfowl. *Sci. Total Environ.* 349, 277–83.
757 doi:10.1016/j.scitotenv.2005.05.021
- 758 Sánchez-Virosta, P., Espín, S., García-Fernández, A.J., Eeva, T., 2015. A review on exposure and
759 effects of arsenic in passerine birds. *Sci. Total Environ.* 512–513, 506–525.
760 doi:10.1016/J.SCITOTENV.2015.01.069
- 761 Scarton, F., 2017. Long-term trend of the waterbird community breeding in a heavily man-modified
762 coastal lagoon: the case of the important bird area ‘‘Lagoon of Venice’’. *J. Coast. Conserv.* 21,
763 35–45. doi:10.1007/s11852-016-0470-8
- 764 Spahn, S.A., Sherry, T.W., 1999. Cadmium and lead exposure associated with reduced growth
765 rates, poorer fledging success of little blue heron chicks (*Egretta caerulea*) in south Louisiana

- 766 wetlands. Arch. Environ. Contam. Toxicol. 37, 377–84.
- 767 Spallholz, J.E., Hoffman, D.J., 2002. Selenium toxicity: cause and effects in aquatic birds. Aquat.
768 Toxicol. 57, 27–37. doi:[https://doi.org/10.1016/S0166-445X\(01\)00268-5](https://doi.org/10.1016/S0166-445X(01)00268-5)
- 769 St. Clair, C.T., Baird, P., Ydenberg, R., Elner, R., Bendell, L.I., 2015. Trace elements in pacific
770 Dunlin (*Calidris alpina pacifica*): patterns of accumulation and concentrations in kidneys and
771 feathers. Ecotoxicology 24, 29–44. doi:10.1007/s10646-014-1352-1
- 772 Tartu, S., Goutte, A., Bustamante, P., Angelier, F., Moe, B., Clément-Chastel, C., Bech, C.,
773 Gabrielsen, G.W., Bustnes, J.O., Chastel, O., 2013. To breed or not to breed: endocrine
774 response to mercury contamination by an Arctic seabird. Biol. Lett. 9, 20130317.
775 doi:10.1098/rsbl.2013.0317
- 776 Taylor, C.E., Cristol, D.A., 2015. Tissue mercury concentrations and survival of Tree swallow
777 embryos, nestlings and young adult females on a contaminated site. Bull. Environ. Contam.
778 Toxicol. 95, 459–464. doi:10.1007/s00128-015-1643-0
- 779 Varian-Ramos, C.W., Swaddle, J.P., Cristol, D.A., 2014. Mercury reduces avian reproductive
780 success and imposes selection: An experimental study with adult- or lifetime-exposure in
781 Zebra finch. PLoS One 9, e95674. doi:10.1371/journal.pone.0095674
- 782 Whitney, M.C., Cristol, D.A., 2017. Impacts of sublethal mercury exposure on birds: A detailed
783 review. Springer, Cham, pp. 113–163. doi:10.1007/398_2017_4
- 784 Zheng, S., Wang, P., Sun, H., Matsiko, J., Hao, Y., Meng, D., Li, Y., Zhang, G., Zhang, Q., Jiang,
785 G., 2018. Tissue distribution and maternal transfer of persistent organic pollutants in Kentish
786 Plovers (*Charadrius alexandrinus*) from Cangzhou Wetland, Bohai Bay, China. Sci. Total
787 Environ. 612, 1105–1113. doi:10.1016/J.SCITOTENV.2017.08.323