

1 Title:

2 A Petri net modeling approach to explore the temporal dynamics of the provision of multiple ecosystem  
3 services.

4

5 Keywords:

6 Ecosystem services interactions; integrated management; social-ecological systems; Venice lagoon; climate  
7 change.

8

9 Abstract:

10 The representation of the temporal dynamics of ecosystem services (ES) is a crucial research frontier in the  
11 field of ES modeling. In fact, most current ES models focus on static ES assessments, that need to be  
12 repeated with different inputs per time step to explore potential changes in ES. Here, we present a new  
13 approach for the dynamic modeling of multiple ES, based on the Petri Net modeling framework. The key  
14 features are: (i) multiple ES are modeled together as a single network, using a social-ecological systems  
15 (SES) perspective; (ii) the model accounts for the interactions occurring among ES, by distinguishing  
16 between the ES whose provision is mediated by some type of human input, which can produce some side-  
17 effects on the system, and those that are generated directly through ecosystem functions and do not  
18 generate side-effects; (iii) the model can reproduce the effects of changing drivers on the elements of the  
19 SES. These features allow to use the model to explore how ES can evolve over time under different “what-  
20 if” scenarios. The importance of considering the ES interactions is tested, showing that failing to include  
21 them in the model remarkably affects the results. Due to its complexity, the model should be used as an  
22 exploratory tool, focusing on the analysis of the general trends of multiple ES provision, rather than on the  
23 generation of quantitative projections. A first conceptual application to the Venice lagoon, Italy, is  
24 presented, in which the trends of 13 different ES are simulated. This application shows the potential of the  
25 model in exploring the development produced by climate change and socio-economic pressures, and the  
26 effects of a set of possible management actions. This modeling approach can contribute to generate new  
27 perspectives on the dynamic modeling of multiple ES and on the integrated management of SES.

28

## 29 1. Introduction

30 Ecosystem services (ES) emerge from the complex interactions occurring between ecosystems and humans,  
31 within the context of interconnected social-ecological systems (SES) (Fischer and Eastwood, 2016; Ostrom,  
32 2009; Reyers et al., 2013). SES are complex adaptive systems characterized by complex processes,  
33 feedbacks and trade-offs which cannot be captured if social and ecological systems are studied separately  
34 (Levin et al., 2013; Liu et al., 2007). Therefore, the study of SES and the ES that they produce requires  
35 system-based methods of analysis that account for their complexity (Bennett et al., 2015; Reyers et al.,  
36 2013).

37 Several ES modeling tools exist (e.g. Boumans et al., 2015; Jackson et al., 2013; Sharp et al., 2014; Tallis and  
38 Polasky, 2009; Villa et al., 2014), that provide useful tools for the assessment of multiple ES and for  
39 generating ES predictions under various scenarios (for a review see Bagstad et al., 2013; Ochoa and Urbina-  
40 Cardona, 2017; Rieb et al., 2017). However, current ES models lack to account for some key elements of  
41 complexity, with respect to three main aspects: space-time ES dynamics, link with human well-being and  
42 the role of technology in enhancing and/or substituting ES (Rieb et al., 2017). With respect to the temporal  
43 dynamics, in fact, most of the current ES modeling approaches focus on the static prediction of the ES  
44 provision, providing a snapshot referred to a single step in time (Rieb et al., 2017). As a result, trends are  
45 often explored by running the models multiple times with different inputs (e.g. Rukundo et al., 2018; Xu et  
46 al., 2018), often based on land use data (e.g. Lautenbach et al., 2011; Stürck et al., 2015), rather than by  
47 modeling the ES dynamics. The global unified metamodel of the biosphere (GUMBO) (Boumans et al.,  
48 2002), and the Multiscale Integrated Models of Ecosystem Services (MIMES) (Boumans et al., 2015) are ES  
49 models based on system dynamics, which are designed to simulate the dynamics of multiple ES. However,  
50 the lack of documentation and methodological support has hindered their application in scientific studies  
51 so far (Ochoa and Urbina-Cardona, 2017).

52 This paper presents a new approach for modeling the temporal dynamics of the provision of multiple ES. It  
53 builds upon the social-ecological viewpoint for ES analysis proposed by Rova and Pranovi (2017), turning it

54 into an dynamic model using Petri nets (Girault and Valk, 2003; Murata, 1989). A previous application of  
55 Petri nets to ES exists (Fongwa et al., 2010), aimed at providing a decision-support system for agro-forest  
56 landscapes, which supported the choice of Petri nets as modeling framework for this work. Petri nets, in  
57 fact, are characterized by a graphical structure that facilitates the communication of the modeling work to  
58 stakeholders, and furthermore they allow the modeler to fully specify the model structure, functions and  
59 parameters, and to represent a variety of different ecological and social processes and interactions. A first,  
60 conceptual application to the Venice lagoon, Italy, is presented, that represents a set of 13 ES provided by  
61 the lagoon SES. The Venice lagoon is an excellent example of complex SES, in which nature and humans  
62 have coexisted for centuries, with a co-evolution which has resulted in profound modifications of both the  
63 lagoon ecosystem and the habits of the local society (D'Alpaos, 2010; Ravera, 2000; Solidoro et al., 2010).  
64 This deep linkage between social and ecological aspects, and the urgent threats related to climate change  
65 have posed the challenge to develop a new ES modelling approach, flexible enough to represent the  
66 peculiar characteristics of the Venice lagoon SES, and capable to dynamically simulate the production of  
67 multiple ES under different scenarios.

68 In particular, this manuscript addresses three main research questions: (1) How can multiple ES be modeled  
69 together accounting for their interactions and dynamics? (2) From a first, explorative application to the  
70 Venice lagoon case study, how might the current drivers of change and climate change pressures affect the  
71 multiple ES delivered by the lagoon? (3) Can we use the model to explore which management actions could  
72 be effective in maintaining the provision of ES over time?

## 73 2. Materials and methods

### 74 2.1. Modeling approach

75 The structure of the model has been developed by making use of the tiered structure of Ostrom's SES  
76 framework (Ostrom, 2009), based on four core subsystems (resource system, resource units, actors and  
77 governance system) and their interactions (McGinnis and Ostrom, 2014; Ostrom, 2009). This allows to include  
78 both ecological and social elements involved in ES' delivery. The model reflects the approach proposed by

79 Rova and Pranovi (2017), which has been translated into the general Petri Net structure shown in Figure 1  
80 (please see below for a brief introduction to Petri nets). The model makes a distinction between ES with  
81 direct and mediated flow types (*sensu* Rova and Pranovi (2017)), that is, it differentiates between ES provided  
82 directly through ecosystem functions, occurring independently of human inputs (direct flow type), and ES  
83 whose provision is mediated by human activities that “use” the resource (mediated flow type). For example,  
84 climate regulation is a direct ES, as it depends e.g. on coastal habitats’ carbon sequestration function (Figure  
85 1A), whereas seafood is a mediated ES because it necessarily depends on fishing activities (Figure 1B).  
86 Activities are performed by actors (e.g. fishermen) and can be regulated by the governance system (e.g. a  
87 fishery management institution). The crucial difference between these two types of ES is that the flow of  
88 direct ES does not consume resources and does not generate negative effects on the system, whereas the  
89 activities of the mediated ES can (and often do) consume the resource units upon which they depend, and,  
90 most of all, can generate negative side-effects on other resources (externalities). For example, fishing  
91 activities can produce negative impacts on coastal habitats, thus affecting the provision of other ES. The  
92 modeling of the activities and their impact on the system is the key characteristic of the present approach,  
93 which allows to represent not only how multiple ES are produced but also the way they interact with each  
94 other.

95 In this work, colored continuous Petri nets are used to model multiple ES. Petri nets are graphical and  
96 mathematical modeling tools, represented as directed, weighted, bipartite graphs (see Murata (1989),  
97 Girault and Valk (2003) and Esparza and Nielsen (1994) for surveys on Petri nets and their properties). They  
98 consist of two kinds of nodes: places, generally representing conditions, items or resources (drawn as  
99 circles), and transitions, generally representing events or (re)actions (drawn as boxes). In this work, places  
100 are used to represent ES, resource systems, actors and governance system (e.g. seafood ES, fish stock,  
101 fishermen and fishery management institution), whereas transitions represent the interactions among the  
102 elements of the SES (i.e. processes, ecosystem functions and activities) (Figure 1).

103 Directed arcs, drawn as arrows, connect nodes of different type, so that transitions have a certain number  
104 of input places (preconditions, items needed for the action) and outputs places (postconditions, items

105 produced). For example, fish and fishermen are inputs for the fishing activity, that generates the seafood  
106 ES. Different types of arcs are used to represent different types of relationships between a transition and its  
107 input places. In particular, normal arcs (drawn as solid arrows) imply that the transition consumes the  
108 resources contained in the input places (e.g. the fishing activity removes fish from the stock); “read arcs”  
109 (drawn as solid lines ending with a circle) imply that the resources in the input place are needed but not  
110 consumed (e.g. the fishing activity requires the fishermen, but does not consume them, or, similarly, in case  
111 of direct ES, the carbon sequestration function depends on habitats but does not consume them);  
112 “modifier arcs” (drawn as dashed arrows) imply that the input places are not needed to enable the  
113 transition but can modify its rate (e.g. a fishing management institution is not a precondition for fishing to  
114 take place, but can modify its rate). For what concerns the negative externalities produced by the activities,  
115 they are represented by weighted arcs (with weights different from one) connecting the impacted  
116 resources with the impacting activity. The weights quantify the magnitude of these side effects, e.g. for the  
117 fishing activity, the amount of habitat consumed per unit of fish caught. In this way, the model can  
118 represent the loss of habitats connected with the fishing activity, and thus the trade-off occurring between  
119 seafood ES and other ES delivered by the impacted habitats.

120 In continuous Petri nets (Heiner et al., 2008), a non-negative real number (called “mark”) is specified for  
121 each model variable, representing its “amount”, e.g. the stock of resources available. The arrangement of  
122 marks over the net (a vector called “marking”) specifies the overall system state. Furthermore, rate  
123 functions, which can be any kind of mathematical function and express the “speed” of the transformation  
124 from input to output places (Heiner et al., 2010), are assigned to all transitions. For example, the rate of the  
125 fishing activity represents the amount of seafood harvested per each time step, calculated as a function of  
126 fish stock, fishermen and governance system. The rate functions are translated and solved as differential  
127 equations when the model simulations are run.

128 Finally, colored Petri nets (Jensen, 1997), were chosen for this work because they allow a compact model  
129 representation. “Colorsets” (sets of one or more colors), which are associated to places, specify, in a tiered-  
130 structure based on the SES framework, the different types of element (e.g. habitats, fauna, actors, etc.)

131 involved in the model. This allows to group and overlay (folding) the portions of the net that represents ES  
132 whose generation involves the same types of elements, resulting in a compact model structure.

133 All the modeling work has been developed using the Petri net tool Snoopy (Heiner et al., 2012; Snoopy,  
134 2017).

## 135 *2.2. Application to the Venice lagoon.*

136 The application to the Venice lagoon, Italy (Figure 2) provides a representation of a set of 13 ES produced  
137 by the lagoon SES (Table 1), and their interactions. The model includes the ES which have been found to be  
138 relevant for the VL in previous studies (Rova et al., 2015; Rova and Pranovi, 2017), and for which a scientific  
139 understanding is currently available .The main effort in the building of the model was put in obtaining a  
140 topology able to catch the multiple ES, their interactions and the cause-effect relationships with drivers of  
141 change, with no ambition of being quantitatively calibrated.

### 142 *2.2.1. Ecosystem services model structure*

143 The workflow starts with the identification of the model variables, which have been organized according to  
144 the tiered structure of the SES framework: based on the core-subsystems of the SES framework (resource  
145 systems and units, actors, governance system) and ES, the types of elements that compose the system have  
146 been specified (colorsets), along with the elements belonging to each of them (colors within each colorset)  
147 (Table 2). The ES have been analyzed and characterized based on (i) the type of ES flow (direct/mediated),  
148 (ii) the resource systems upon which the ES depend, and (iii) the generation of negative externalities,  
149 according to the logical flow depicted in Figure 3. The ES with similar characteristics have been grouped  
150 together, resulting in six ES “topological” groups, which share a similar net topology (Figure 3). Therefore,  
151 by taking advantage of the features of colored Petri nets, a “folded” net structure has been developed for  
152 each “topological” group, resulting in six folded ES subnets (Figure 3). Each folded subnet is based on the  
153 general structure of Figure 1, but incorporates the specific features of each “topological”group. The folded  
154 subnets are a compact way to graphically represent the net structure of the ES belonging to each  
155 “topological”group, as if they were stacked together. Within each subnet, each ES has a specific

156 combination of elements involved (the colors of the places' colorsets) and specific parameters for the  
157 transitions' rate functions. The "unfolding" of the net returns the topology for each ES, which is  
158 summarized in Table 3 and more extensively described in Appendix A.

159 In general, regulating ES follow the general structure developed for the direct flow type ES (Figure 1A),  
160 whereas provisioning and cultural ES follow that of a mediated flow type ES (Figure 1B). Then, each subnet  
161 presents some variations that account for the specific characteristics of the ES "topological" group.  
162 Provisioning 1 ES' subnet (Figure 4B), which refers to low impact fishing and hunting activities, does not  
163 include places impacted through negative externalities. In the model, these activities were assumed to  
164 produce no externalities because the side effects that they produce on other resources (other than the  
165 exploited ones) are extremely low if compared to the habitats' degradation, enhanced channels' siltation  
166 and disturbance to the nursery function that are instead caused by the mechanical harvesting activities  
167 involved in provisioning 2 ES (clam harvesting, Figure 4C) (cfr. Pranovi et al., 2004, 2003). Cultural 1 and 2  
168 ES (Figure 4D-E), have been modeled to be dependent on habitats, heritage and channels, which reflect  
169 natural attractiveness, cultural attractiveness and accessibility, respectively. The difference between these  
170 two "topological" groups concerns the negative externalities. Cultural 2 ES (tourism) produces severe side  
171 effects related to the intensive navigation activities through which visiting occurs, that cause degradation of  
172 habitats and enhanced channels siltation. Cultural 1 ES are instead characterized by slower navigation  
173 modes (rowing and sailing boats, or slow motorboats used for educational excursions) whose negative  
174 impacts can be considered negligible compared to tourism. Cultural 3 ES (navigation, Figure 4F) depends  
175 mainly on the presence of channels and, similarly to tourism, causes channels' siltation and habitats'  
176 degradation.

177 For what concerns the graphical representation of the model, please note that Figures 4, 5 and 6 (which are  
178 described in this section, section 2.2.2 and 2.2.3 respectively) compose together the overall model  
179 structure, which has been split in different portions for visualization purposes; the nodes in grey ("logical  
180 nodes") appear multiple times as graphical copies of a single node, logically identical.



181 *2.2.2. Underpinning ecological and social processes*

182 The model includes, with a certain degree of simplification, the ecological processes and the anthropic  
183 interventions upon which the presence of resource units depend, and can simulate the social trends of  
184 actors' populations. Due to its complexity and variety of variables and processes included, the model  
185 provides a simplified representation of ecological and social processes: an effort was made to design a  
186 model structure that applies a relatively homogeneous degree of simplification to all processes, to avoid  
187 having an imbalance between the detailed representation of some aspects and simplification of others. This  
188 section describes the folded net structure representing these processes. A more detailed description is  
189 provided in Appendix A.

190 Habitats are generated through ecological processes that depend on the extent of each habitat and are  
191 modulated by fauna (target fish species and birds) (Figure 5A). This modulation reflects the feedback of  
192 higher levels of the tropic network on habitats. Furthermore, habitats can be the object of management  
193 actions controlled by the governance system, aimed at their maintenance and/or reconstruction. In  
194 addition, the model accounts for the positive effect of the environmental sensibilization deriving from the  
195 information for cognitive development and tradition ES. This reflects the environmental friendly behavior  
196 of the people that have been exposed to these ES.

197 Channels' presence and navigability are determined by two factors in the model (Figure 5B). The first is self-  
198 regulation capacity, that represents the effects of channels' hydrodynamics on sedimentation. It is  
199 influenced by the erosion prevention 1 and 2 ES, which contribute to prevent siltation. The second factor  
200 are channel dredging activities, regulated by the channel dredging governance system.

201 The abundance of fauna depends on population growth (Figure 5C). Growth depends on the abundance of  
202 the fauna resource units, and is modulated by the lifecycle maintenance ES, reflecting the key role played  
203 by the spawning, nursery and nesting functions for the maintenance of these resources.

204 An actors' growth transition (Figure 5D) allows for the specification of social trends regarding actors, in  
205 particular residents and tourists.

206 The model does not include processes that “produce” cultural heritage and traditions. These resources  
207 derive from past states of the SES, and result from the long-term coevolution between society and  
208 ecosystem. These processes are not modeled as they have a time scale far longer than that of the other  
209 processes considered (please refer to Rova and Pranovi (2017) for a more thorough discussion of these  
210 aspects).

### 211 *2.2.3. Effects of drivers of change*

212 The model simulates the effects of the relative sea level rise (RSLR) and temperature increase driven by  
213 climate change, and the effects of the mobile barriers at the lagoon inlets (MOSE system (Consorzio  
214 Venezia Nuova, 2018)), which are expected to be completed in 2019 in order to defend Venice from  
215 flooding (Figure 6).

216 RSLR (Figure 6A) has been assumed to produce three major effects on the lagoon SES: a negative impact on  
217 salt marshes and bare (intertidal) habitats (Marani et al., 2007; Rizzetto and Tosi, 2011), and seagrasses  
218 (Saunders et al., 2013); a negative impact on residents and an effect on cultural heritage that is initially  
219 positive (increased attractivity) and then negative, as the RSLR increases. The negative effects on residents  
220 and cultural heritage are related to the flooding of urban areas, which increases with increasing water level,  
221 as shown by the altimetric charts of the historical center of Venice (Comune di Venezia, 2018). The  
222 frequency and severity of high tides is expected to increase with RSLR (Carbognin et al., 2010), thus  
223 exacerbating the flooding events. The initial positive effect on cultural heritage has been assumed here to  
224 account for the increased tourist attractivity of the flooded urban areas.

225 The MOSE system consists of a system of gates, installed on the bottom of the three lagoon’s inlets, which  
226 will be raised during high tide events (>110 cm with respect to Punta della Salute tide gauge), temporary  
227 separating the lagoon from the sea. The frequency of high tides is expected to increase with RSLR, and so  
228 the frequency of the MOSE closures (Carbognin et al., 2010; Umgiesser and Matticchio, 2006). In this  
229 model, the yearly frequency of closures is calculated as a function of RSLR, according to the trends  
230 estimated by Carbognin et al. (2010) (Figure 6A). It has been assumed to produce both social and ecological

231 effects: on the one hand, it balances the effects produced by RSLR on residents and cultural heritage, and  
232 on the other hand, because of the modified lagoon-sea exchanges related to the inlets' closure, it  
233 negatively affects submerged habitats, lifecycle maintenance and channels' self-regulation capacity.

234 The effects of temperature increase on habitats and fauna (Figure 6B) have been modeled according to the  
235 following assumptions. Target fish species have been assumed not to change at low levels of temperature  
236 increase (simulating the effects of species substitution), and to be negatively affected at higher levels  
237 (Pranovi et al., 2013). Seagrasses have been assumed to be positively affected at low levels of temperature  
238 increase, and negatively affected at higher levels of temperature increase, which seem to reduce  
239 seagrasses growth when occurring concurrently with a reduced light availability, as that caused by RSLR  
240 (Bulthuis, 1987). A similar behavior has been assumed for clams, as high values of temperature increase  
241 seem to become a stress factor for this species (Munari et al., 2011; Velez et al., 2017).

242 A more detailed description of the modeling of the effects of these drivers is provided in Appendix A.

#### 243 *2.2.4. Rate functions and parameters*

244 The rate functions of all transitions are reported in Table A1 (Appendix A). Wherever possible, functions  
245 widely used in ecology (e.g. logistic population growth) were used, where not possible, the functions reflect  
246 the authors' hypothesis on the modeled processes. The model's initial conditions, functions' parameters  
247 and arc weights (Tables A2, A3, A4 of the Appendix A, respectively) are built up to reproduce the realistic  
248 proportions between the modeled variables and the relative magnitude of the processes occurring in the  
249 lagoon system. Overall, the model setup was tuned to represent an ideal configuration of the Venice lagoon  
250 SES in which all variables are in steady-state. The steady state is a hypothetical perfectly equilibrium  
251 situation, in which all variables are constant over time: no growth function is specified for actors, no  
252 climate change pressure occurs, and resources' consumption perfectly balances their generation rate. As a  
253 result, the ES provision is constant over time too. Moving from this condition, the behavior of the model  
254 was tested by performing a set of simulations in which all the model parameters were changed one at a  
255 time by  $\pm 10\%$  and  $\pm 25\%$ . The analysis was repeated iteratively while tuning the parameters, in a sort of

256 sensitivity analysis, until a satisfactory model behavior was obtained, that broadly reflected the processes  
257 and variables' interactions observed in the lagoon. The results of these simulations, relative to the final  
258 version of the model, are reported in Appendix B. Furthermore, we have tested the sensitivity of the model  
259 with a limited number of combined variations of model variables, with a focus on the governance system's  
260 variables (about 150 combinations, including combination of two, three, four and five variables, with  
261 positive and negative variations). Under these conditions the model showed an overall consistent behavior,  
262 with no ecological nonsenses, and a sensitivity in the same order of magnitude than that obtained with  
263 single variations.

#### 264 *2.2.5. Scenarios*

- 265 - A **Business as usual (BAU)** scenario, that features the deviations from the steady state that  
266 characterize the current situation of the Venice lagoon. These deviations are: (i) increasing tourists,  
267 (ii) decreasing residents, (iii) unbalanced consumption of salt marshes, (vi) increasing seagrasses.  
268 These deviations take place simultaneously and for the entire simulation period. The corresponding  
269 variations in the input parameters are specified in Tables A2 and A3 (Appendix A).
- 270 - Three **Business as usual + Climate change (CC)** scenarios, that incorporate climate change  
271 pressures into the BAU scenario. The simulations include tree RSLR scenarios (15 cm, 25 cm and 50  
272 cm RSLR by the end of the 21<sup>st</sup> century) combined to one temperature scenario (1°C temperature  
273 increase by the end of the 21<sup>st</sup> century), resulting in three CC scenarios (named CC\_15, CC\_25,  
274 CC\_50, respectively).
- 275 - Three **Business as usual + Climate change + MOSE (CC\_MOSE)** scenarios, in which the functioning  
276 of the MOSE system has been combined with the three CC scenarios (resulting in three CC\_MOSE  
277 scenarios named CC\_MOSE\_15, CC\_MOSE\_25, CC\_MOSE\_50 respectively).
- 278 - **Additional management options** scenarios, that feature additional management strategies tested  
279 under BAU and CC\_MOSE scenarios. These strategies include single and combined variations of all  
280 governance systems' management fields (except for MOSE, which is already active under MOSE\_CC  
281 scenarios), aimed at exploring if and how it is possible to balance the negative effects of these

282 scenarios on ES. For the management fields related to the mediated ES' activities (tourism,  
283 navigation, artisanal fishing, recreational fishing, clam harvesting and hunting, which directly  
284 modulate the respective activities' rates), a variation of -50% has been used. For the management  
285 fields related to habitats maintenance and channels dredging, which represent the yearly  
286 maintenance rate expressed as proportion of the resources' initial condition, a variation of +1% has  
287 been used.

288 All simulations have been run until the end of the century.

#### 289 *2.2.6. Aggregated indicators of ES provision*

290 The model outputs illuminate the trends of all its variables over time. To summarize and compare the effects  
291 of the various scenarios on the multiple ES, two aggregated indicators have been developed and computed  
292 based on the ES state at the end of the century:

- 293 - Sum of direct ES' percentage variations with respect to initial conditions ( **$\Delta Dir$** );
- 294 - Sum of mediated ES' percentage variations, excluding tourism, with respect to initial conditions  
295 ( **$\Delta Med-T$** ).

296 Tourism ES was not included in  $\Delta Med-T$  because, being the major driver of change in the BAU scenario, it  
297 was expected to show a distinct trend. Therefore, its variation has been considered separately.

#### 298 *2.2.7. Testing the effects of excluding the interactions among ecosystem services*

299 To sum up, the multiple ES included in the model interact, either directly or indirectly, in the following  
300 ways:

- 301 a) consumption of the same resource units (i.e. artisanal and recreational fishing activities insisting on  
302 the target fish species);
- 303 b) negative effects generated by some of the activities of the mediated ES (i.e. negative effects of  
304 clam harvesting, tourism and navigation ES);

- 305 c) positive effects of some ES on the resource systems (i.e. the environmental sensibilization deriving  
306 from the information for cognitive development and tradition ES, and the effect of erosion  
307 prevention ES on channels' self-regulation);
- 308 d) ecological feedbacks (i.e. fauna influencing the habitats' processes, and lifecycle maintenance ES  
309 influencing the growth of fauna).

310 The importance of having these interactions included in the model was tested by analyzing the effects that  
311 their exclusion has on the model results. To do so, three additional model configurations were created,  
312 which neglect the ES interactions partially or completely:

- 313 - a configuration without the positive and negative side effects produced by ES (points (b) and (c)  
314 above) ("NO\_ES\_sideEffects"). This configuration represents a model that mainly ignores the  
315 interactions deriving from "social" aspects of ES delivery (i.e. the consequences of human  
316 activities);
- 317 - a configuration without the ecological feedbacks (point (d) above) ("NO\_EcoFeedbacks"). This  
318 configuration, on the other hand, represents a model that ignores the interactions deriving from  
319 the "ecological" aspects of ES delivery (i.e. the feedbacks between ecological elements);
- 320 - a configuration without both ("NO\_ALL").

321 The first source of interaction listed above (point (a)) could not be excluded because it would require  
322 eliminating one of the two fishing ES. For details on the setup of these configurations, please refer to  
323 Appendix A. The BAU and CC\_MOSE scenarios were run with each of these configurations to compare the  
324 different outcomes.

## 325 **3. Results**

### 326 *3.1. Business as usual, climate change and MOSE scenarios*

327 Figure 7 shows the relative variation over time of the 13 ES considered in this study, under the BAU  
328 scenario. The massive loss of ES indicates that the BAU is an unsustainable scenario, even without

329 considering the potential effects of climate change. Management actions are thus necessary to prevent the  
330 decline of ES over time. A trade off can be observed between tourism ES, whose marked increase is driven  
331 by the growing number of tourists assumed as BAU's major driver, and all the other ES, which are instead  
332 characterized by a general declining trend, except for erosion prevention 2. This trend shows that the  
333 model is capable to represent the feedbacks of socio-economic drivers (increase of the number of visitors  
334 and decrease of residents) on the lagoon ecosystem and on the ES it produces. The aggregated indicators  
335  $\Delta Dir$  and  $\Delta Med-T$ , and tourism variation (Figure 8) synthetically represent these trends.

336 The effects of CC scenarios (combination of RSLR, 15, 25 and 50 cm, and 1°C temperature increase) and  
337 CC\_MOSE scenarios on the overall ES provision at the end of the 21<sup>st</sup> century are compared using the  
338 aggregated indicators  $\Delta Dir$  and  $\Delta Med-T$ , and tourism variation (Figure 8). All these indicators are  
339 progressively reduced under more extreme CC scenarios. The functioning of MOSE does not change this  
340 overall trend, but produces different effects on the three indicators: (i) it does not offset the loss of direct  
341 ES, but rather tends to intensify the reduction of  $\Delta Dir$  in the more extreme CC scenarios; (ii) it has a positive  
342 effect on  $\Delta Med-T$  with respect to CC\_15 and CC\_25 scenarios, but fails to produce an improvement with  
343 respect to CC\_50; (iii) it has a positive impact on tourism in all cases, this effect becoming greater under  
344 more extreme scenarios. In any case, the MOSE system alone is not sufficient to prevent the effects of  
345 climate change on the multiple ES, and thus it requires to be combined with additional management  
346 options (the variation of ES over time under CC\_MOSE scenarios is shown in Figures C1, C2 and C3 of  
347 Appendix C).

### 348 *3.2. Additional management options*

349 Single additional management options have been tested under BAU and CC\_MOSE scenarios, and their  
350 effectiveness has been evaluated with respect to the values assumed by the  $\Delta Dir$  and  $\Delta Med-T$  indicators.  
351 The target for considering these interventions successful is the compensation of the reduction of these  
352 indicators with respect to the initial conditions. The two aggregated indicators have been given priority  
353 with respect to tourism's variation as long as the latter does not show a decrease with respect to the initial  
354 conditions.

355 The single management options have been ranked based on their effectiveness with respect to each  
356 indicator (Table 4). The ranking is nearly the same in all scenarios. Maintenance of seagrasses produces the  
357 greatest effects in all cases, however, there is no case in which a single option can be effective in balancing  
358 both indicators. The lack of effectiveness of sectorial management points out the need to enforce  
359 management actions that operate at ecosystem level, combining different options together. To account for  
360 this, the following combinations of two, three and four management options have been tested in the  
361 model, designed based on the top three options of the rankings shown in Table 4.

362 Combinations of two:

- 363 • Seagrass maintenance & Salt marsh maintenance
- 364 • Seagrass maintenance & Tourism
- 365 • Seagrass maintenance & Benthic diatoms maintenance

366 Combinations of three:

- 367 • Seagrass maintenance & Salt marsh maintenance & Benthic diatoms maintenance
- 368 • Seagrass maintenance & Tourism & Benthic diatoms maintenance
- 369 • Seagrass maintenance & Tourism & Salt marsh maintenance

370 Combination of four:

- 371 • Seagrass maintenance & Tourism & Salt marsh maintenance & Benthic diatoms maintenance

372 Figure 9 summarizes the effects produced by these combinations under the four scenarios, with respect to  
373 the two aggregated indicators. In the case of combinations, the most effective solution(s) can be identified  
374 as that(those) meeting the target (counteracting the ES reduction with respect to initial conditions) with  
375 the fewest management options involved. Concerning  $\Delta Dir$ , combinations of two options are effective up to  
376 CC\_MOSE\_25, but fail to balance the loss of direct ES in CC\_MOSE\_50, for which a combination of three  
377 options is needed. Regarding  $\Delta Med-T$ , the management options seem less effective than in case of direct  
378 ES. The combinations of two options are insufficient also in case of CC\_MOSE\_25, for which only seagrass



379 maintenance + tourism is effective. For CC\_MOSE\_50, the combination of seagrasses maintenance +  
380 tourism + benthic diatoms maintenance is the only one that fully balances this indicator, and seagrasses  
381 maintenance + tourism + salt marshes maintenance is almost effective with a reduction of about -1%.  
382 Overall, the target can be met for both indicators under all scenarios only if combinations of three  
383 management options are enforced, which combine the maintenance of seagrass and either salt marshes or  
384 diatoms habitats with the reduction of tourism.

### 385 *3.3. Effects of excluding the interactions among ecosystem services*

386 If the interactions among ES are excluded from the model, we obtain a situation in which the multiple ES  
387 are isolated from each other. The consequences of this exclusion are visible by comparing the results  
388 obtained from the complete model with those obtained from the three configurations in which the ES  
389 interactions were removed partially or completely (Table 5).

390 Looking at the BAU scenario, it appears that the lack of consideration of the ES interactions results in  
391 markedly different trends for most of the ES. In particular, the negative trends of many regulating ES are  
392 not captured if the ES side effects are not considered (NO\_ES\_sideEffects configuration). With this  
393 configuration, these ES even show an overall positive trend that is in net contrast with the negative one  
394 revealed by the complete model ( $\Delta$ Dir aggregated indicator). Additionally, both the configurations lacking  
395 either the ecological feedbacks or the ES side effects (NO\_EcoFeedbacks and NO\_ES\_sideEffects) fail to  
396 capture the negative trend of the provisioning ES, that is related to the deterioration of the ecological  
397 conditions occurring under this scenario. By comparing the NO\_EcoFeedbacks and NO\_ES\_sideEffects  
398 configurations, it appears that the second deviates from the complete model more than the first,  
399 suggesting that the consideration of social aspects (such as human activities and their side effects) is crucial  
400 to understand the system behavior. Overall, neglecting the interactions among ES (NO\_ALL configuration)  
401 would lead to a radically different interpretation of the BAU scenario, which could be misleadingly thought  
402 to have relatively acceptable consequences for the multiple ES provided by the lagoon.

403 The effect of excluding the ES interactions is less pronounced under the CC\_MOSE scenarios. This was  
404 expected, as these scenarios produce direct impacts on all the resource systems, and thus directly affect all  
405 the ES, whose resulting trends can be broadly detected also by a model that considers them separately.  
406 However, it should be noted that the exclusion of the ecological feedbacks (NO\_EcoFeedbacks  
407 configuration) leads to quite different results for the provisioning ES: these ES are only very marginally  
408 reduced, and their negative trend does not increase with more severe scenarios, as instead indicated by the  
409 complete model. This suggests that failing to include the ecological feedbacks leads to a model that is not  
410 fully capable to capture the increasingly severe consequences of the drivers of change. Additionally,  
411 differently from what observed under the BAU scenario, in this case the model results seem to be more  
412 sensitive to the lack of ecological feedbacks, with respect to the lack of ES side effects.

413 Overall, the simulations under the four scenarios show that the results are remarkably different if the ES  
414 interactions are neglected, and in particular, that the interactions deriving from both social and ecological  
415 aspects are of crucial importance for understanding the potential effects of drivers of change (and  
416 management actions) on the system.

## 417 4. Discussion

### 418 4.1. Modeling approach

419 Within the vast panorama of ES models, the dynamic representation of ES has been identified as one of the  
420 crucial research frontiers in ES modeling research (Bennett et al., 2015; Rau et al., 2018; Rieb et al., 2017).  
421 To overcome these limitations, new tools are needed that are capable to simulate, in a dynamic way, the  
422 mechanisms that produce the relationships between ES (Bennett et al., 2015; Rieb et al., 2017), i.e.  
423 interactions among ES and effects of drivers on multiple ES (Bennett et al., 2009; Spake et al., 2017). This is  
424 the direction in which the innovative elements of our modeling approach are going.

425 First, the model includes the social and ecological elements involved in the provision of multiple ES, which  
426 are selected and organized based on the SES framework (McGinnis and Ostrom, 2014; Ostrom, 2009). The  
427 SES framework helps in the identification of variables and processes that are relevant for the analysis,

428 which is the first and very challenging step as it requires simplifications and abstractions to be made  
429 (Schlüter et al., 2014). According to Bennett et al. (2009), an integrated social-ecological approach is the  
430 basis for a better understanding of ES relationships. On these regards, our results show that, from a  
431 modeling perspective, the inclusion of ES interactions deriving from both a social and ecological perspective  
432 is crucial for capturing the ES trends caused by different drivers of change. On the one hand, this underlines  
433 the limits of the modeling tools that consider the ES separately, such as the widely used INVEST (Sharp et  
434 al., 2014; Tallis and Polasky, 2009), that consists in a suite of models, each of which assesses a single ES. On  
435 the other hand, it highlights the need to further develop modeling tools that explicitly incorporate the  
436 interactions among ES. Adopting a social-ecological perspective from the very first steps of model  
437 development is crucial on these regards, as it facilitates the recognition of the interactions among ES, and  
438 subsequently, their implementation in the model. In our work, the social-ecological viewpoint proposed by  
439 Rova and Pranovi (2017), and in particular the distinction between direct and mediated ES, provided a  
440 useful baseline for the identification of the different ways in which ES interact, and their incorporation in  
441 the model.

442 Second, the model is structured as a single network of multiple ES, that emerge from the dynamic  
443 interactions (processes, functions, activities) occurring between the elements of the SES. The bipartite  
444 structure of Petri nets is well suited for this scope, as it alternates places (representing the different  
445 elements of the system) with transitions (representing the interactions between these elements). This  
446 network of ES behaves dynamically according to the rate functions that are associated to the transitions.  
447 The definition of rate functions is a very challenging step, as it requires a substantial simplification of  
448 complex processes and makes explicit the assumptions about the causal relationships between the  
449 variables involved (Schlüter et al., 2014). Finally, drivers of change, such as climate change and increasing  
450 tourism, act upon this model structure by producing changes in the SES resources and actors, which in turn  
451 generate the dynamic response of the whole set of interacting ES. In this way, the model captures both  
452 types of mechanisms that, according to Bennett et al. (2009), produce the relationships between ES, that is,

453 interactions among ES and effects of drivers on multiple ES, and can thus represent the trends of multiple  
454 ES over time.

455 In addition, the dynamic features of the model allow to simulate the effects of management actions on the  
456 system. The evaluation of these actions requires the definition of objectives and performance measures  
457 (Martinez-Harms et al., 2015), which can be calculated based on the model outputs. These measures can be  
458 used to assess the improvements generated by different management options, and thus to prioritize the  
459 actions based on their effectiveness. Despite its potential usefulness for decision making, prioritization of  
460 management actions is still poorly addressed by ES studies (Martinez-Harms et al., 2015). In this work, the  
461 sum of the variations of direct and mediated ES (*sensu* Rova and Pranovi, 2017) were used to evaluate the  
462 performance of the management options, the objective being the compensation of the negative effects of  
463 BAU and CC scenarios on these indicators. The distinction between direct and mediated ES is used here to  
464 keep track of the trends of ES that spontaneously arise from ecosystem functions and do not generate  
465 negative effects (direct ES), and of ES that could produce side effects due to the human inputs involved  
466 (mediated ES). The sum is indeed a very basic way of aggregating multiple ES, as all ES are considered to  
467 have the same importance within each indicator, but represents a first step of analysis.

468 A major limitation of the current model application is the related uncertainty. The uncertainty of our model  
469 mainly lies in the aggregation of variables and in the simplification of the represented processes. This is a  
470 consequence of the focus on multiple ES and their interactions, that increases the overall complexity of the  
471 model. This aspect was addressed by structuring the model in a way that all the processes are characterized  
472 by a similar degree of simplification and by repeatedly checking for an overall consistent model behavior  
473 during model development. Although the model is not calibrated, the sensitivity analysis shows overall  
474 ecologically sound results, with no illogical responses and a relatively low sensitivity to variations of input  
475 data. As a result, the model can be considered quite reliable in the representation of the broad trends  
476 produced by the drivers of change and management options, but should not be expected to provide  
477 quantitative ES predictions. Therefore, the model should be intended as an exploratory modelling tool,  
478 focused on understanding the general system's behavior and trends under different "what-if" scenarios.

479 4.2. Case study application

480 The application to the Venice lagoon case study provides an example of the potential of this tool to  
481 investigate future trends of multiple ES, and to evaluate and prioritize potential management options. Four  
482 main take-home messages emerge from this application:

- 483 1) **The BAU scenario is unsustainable.** The increasing tourism pressure, combined with the decline of  
484 residents and the progressive salt marsh degradation result in a decreasing trend of most  
485 regulating, provisioning and cultural ES. Climate change then acts making these trends more  
486 severe, exacerbating a situation which is already compromised. Therefore, management strategies  
487 cannot focus only on climate change adaptation but need to address, at the same time, the  
488 negative trends that are occurring under the BAU conditions.
- 489 2) **The complex situation requires an integrated management approach.** The model outcomes call  
490 for a holistic management of environmental resources, or “ecosystem approach”, as defined in the  
491 UN Convention of Biological Diversity (see also Borja et al., 2016; Elliott, 2014, 2011), that is, a  
492 management approach that integrates management actions in diverse sectors for a common aim:  
493 maintaining the functioning of the system and the benefits it delivers to society. In fact, as the  
494 model shows, none of the management options tested, individually, is capable to counterbalance  
495 the negative trends observed in any of the scenarios. When multiple ES are modeled  
496 simultaneously, and combined to set management targets, the ineffectiveness of sectorial  
497 management is powerfully highlighted, and, in particular, the combination of multiple management  
498 actions emerges as the only way to balance the negative ES trends in the modeled system.
- 499 3) **Habitats’ conservation and restoration is of primary importance for the provision of multiple ES**  
500 **in the lagoon SES.** Among the additional management options having, individually, the greater  
501 effects, it appears that those targeted to habitat maintenance are promising better outcomes than  
502 those aimed at limiting the environmental pressures insisting on the system, except for tourism  
503 control. This suggests that the ecological elements of the systems are crucial for maintenance of

504 the ecological processes and functions, and for the delivery not only of regulating ES but also of  
505 provisioning and cultural ones.

506 4) **Multiple management options are needed that combine different types of intervention, to be**  
507 **enforced now.** If on the one hand the MOSE system plays a crucial role in maintaining cultural  
508 heritage and tourism in the face of climate change, and has a generally positive effect on the other  
509 mediated ES, on the other hand, it seems to exacerbate the decline of regulating ES, thus requiring  
510 to be combined with other interventions. The type of combination required depends on the  
511 scenario: the more severe the scenarios tested, the more complex the set of management options  
512 needed to offset the negative effects on ES. Considering the uncertainties on how climate change  
513 will evolve, the precautionary principle should be applied, and thus the management solutions that  
514 are effective in the worse scenario should be preferred. It should be noted that, although scenarios  
515 have a time span of decades, the implementation of management actions should start now, to  
516 gradually contribute to make the system more resilient, in the face of potential extreme scenarios.  
517 From the outcomes of the model, the most effective outcomes are obtained through conservation  
518 and restoration of crucial habitats (seagrasses and salt marshes or benthic diatoms), combined with  
519 a reduction of tourism. Tourism indeed plays a controversial role, being on the one hand the main  
520 economic engine of the area, and on the other, a major pressure on the other ES. If maintaining the  
521 provision of multiple ES over time is taken as management priority, it appears necessary to enforce  
522 some control over tourism to balance the loss of other ES under CC scenarios.

523 An interesting field of application of the tool here proposed could be the implementation of the Ecosystem  
524 Approach to transitional water management. In particular, at present, a challenging issue is represented by  
525 the implementation of the Water Framework Directive 2000/60/EC (Voulvoulis et al., 2017). Indeed, it is  
526 not completely clear how to pass from the monitoring of the ecological status, based on biological quality  
527 elements, to the implementation of efficient management strategies to recovery from bad/scarce  
528 conditions. In this context, the management of multiple ES, which depend on the ecological status but also  
529 produce feedbacks on it, supported by a modeling tool capable to capture these feedbacks, could provide a

530 new perspective for shifting from monitoring to implementation. This could be particularly helpful in highly  
531 co-evolved environments, as the Venice lagoon, allowing to produce simulations about possible effects of  
532 different management options.

## 533 5. Conclusions

534 This paper presents a new approach for the dynamic modelling of multiple ES provision, developed using  
535 the Petri net modeling framework. Three key characteristics of the model are of crucial importance for the  
536 representation of multiple ES' dynamics:

- 537 1) the model is structured as a single, complex network that provides a joint representation of the  
538 different ES provided by the system. The bipartite structure of Petri nets, that alternates places  
539 (elements of the system) and transitions (processes, functions, activities) proved to be well suited  
540 for this scope;
- 541 2) the SES perspective plays a crucial role for the model development, for the identification of the  
542 social and ecological elements and processes involved in the provision of the different ES, and for  
543 the identification of the different ways in which these ES interact. In this work, the SES viewpoint by  
544 Rova and Pranovi (2017), and in particular, distinction between direct and mediated ES, has  
545 provided a good foundation for the representation of interactions among ES. Failing to include the  
546 ES interactions in the model remarkably affects the results;
- 547 3) the model's structure can be customized to include the effects of drivers of change on ES. In the  
548 case study application, the core structure of the model, that represents the multiple ES, has been  
549 expanded to incorporate the potential effects produced by different drivers of change on the SES  
550 resources and actors, which are then reflected by changes in ES provision.

551 The first explorative application to the Venice lagoon case study suggests that most ES are declining under  
552 the BAU and CC scenarios, with a major trade-off between tourism and the other ES. The functioning of the  
553 MOSE system does not seem to be sufficient to compensate this decline, and requires to be combined with  
554 other interventions, among which those aimed at habitats' conservation and restoration seem to be the

555 most effective. The major advantage of a model that jointly represents multiple ES is that it can be used to  
556 simulate the effects of very different management actions on the whole set of regulating, provisioning and  
557 cultural ES. Although being less accurate than discipline-specific models, it considers a wide range of direct  
558 and indirect implications that would not emerge from models focused on single ES, and can thus be a  
559 precious support for the definition of integrated management strategies.

560 This first version of the model leaves the floor open to several improvements and further steps. First, the  
561 “ES use” step of the general structure in Figure 1 could be used to model the ES demand by stakeholders,  
562 which is indeed another crucial frontier for ES models (Rieb et al., 2017). This would allow to investigate the  
563 ES synergies and trade-offs that are related to their use, e.g. concurring or conflicting use (Mouchet et al.,  
564 2014). Second, concerning the Venice lagoon case study, the application presented here could be upgraded  
565 to a numerically more realistic model, possibly moving towards a more operational tool. As data about  
566 several input variables and parameters are lacking, the model should be fed with a combination of available  
567 data and expert-based inputs. Third, the model could be used to prioritize management options with  
568 respect to more detailed management targets. More specific targets could imply a prioritization of some ES  
569 over others and/or the definition of specific thresholds of ES provision. This could be obtained from a  
570 deeper SES analysis that connects ES with specific dimensions of human well-being (Reyers et al., 2013),  
571 and/or from the collection of stakeholder preferences (Martinez-Harms et al., 2015). Overall, although still  
572 in its development phase, this modeling approach can hopefully contribute to generate new perspectives  
573 for the dynamic modeling of ES and can be the starting point for more advanced applications aimed at  
574 actively supporting the integrated management of social-ecological systems.

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## 578 References

- 579 Bagstad, K.J., Semmens, D.J., Waage, S., Winthrop, R., 2013. A comparative assessment of decision-support tools for  
580 ecosystem services quantification and valuation. *Ecosyst. Serv.* 5, 27–39.  
581 doi:<http://dx.doi.org/10.1016/j.ecoser.2013.07.004>
- 582 Bennett, E.M., Cramer, W., Begossi, A., Cundill, G., Díaz, S., Egoh, B.N., Geijzendorffer, I.R., Krug, C.B., Lavorel, S.,



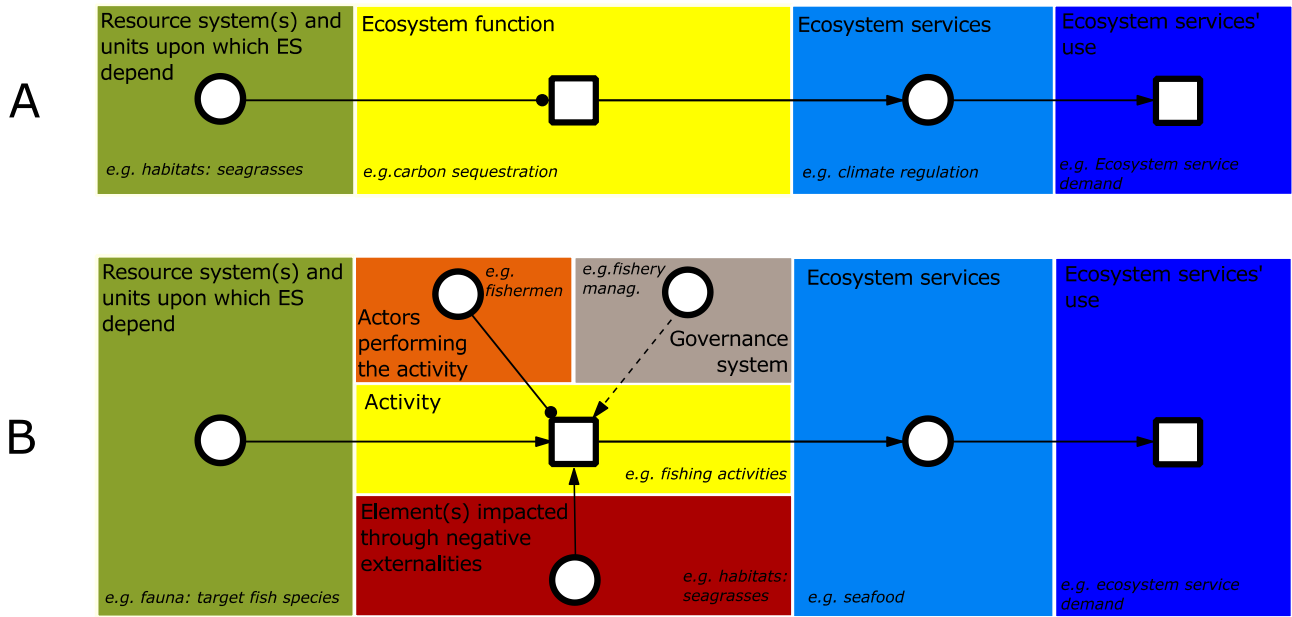
- 583 Lazos, E., Lebel, L., Martín-López, B., Meyfroidt, P., Mooney, H.A., Nel, J.L., Pascual, U., Payet, K., Harguindeguy,  
584 N.P., Peterson, G.D., Prieur-Richard, A.-H., Reyers, B., Roebeling, P., Seppelt, R., Solan, M., Tschakert, P.,  
585 Tschardtke, T., Turner, B., Verburg, P.H., Viglizzo, E.F., White, P.C., Woodward, G., 2015. Linking biodiversity,  
586 ecosystem services, and human well-being: three challenges for designing research for sustainability. *Curr. Opin.*  
587 *Environ. Sustain.* 14, 76–85. doi:10.1016/j.cosust.2015.03.007
- 588 Bennett, E.M., Peterson, G.D., Gordon, L.J., 2009. Understanding relationships among multiple ecosystem services.  
589 *Ecol. Lett.* 12, 1394–1404. doi:10.1111/j.1461-0248.2009.01387.x
- 590 Borja, A., Elliott, M., Andersen, J.H., Berg, T., Carstensen, J., Halpern, B.S., Heiskanen, A.-S., Korpinen, S., Lowndes,  
591 J.S.S., Martin, G., Rodriguez-Ezpeleta, N., 2016. Overview of Integrative Assessment of Marine Systems: The  
592 Ecosystem Approach in Practice. *Front. Mar. Sci.* 3, 1–20. doi:10.3389/fmars.2016.00020
- 593 Boumans, R., Costanza, R., Farley, J., Wilson, M.A., Portela, R., Rotmans, J., Villa, F., Grasso, M., 2002. Modeling the  
594 dynamics of the integrated earth system and the value of global ecosystem services using the GUMBO model.  
595 *Ecol. Econ.* 41, 529–560. doi:10.1016/s0921-8009(02)00098-8
- 596 Boumans, R., Roman, J., Altman, I., Kaufman, L., 2015. The Multiscale Integrated Model of Ecosystem Services  
597 (MIMES): Simulating the interactions of coupled human and natural systems. *Ecosyst. Serv.* 12, 30–41.  
598 doi:http://dx.doi.org/10.1016/j.ecoser.2015.01.004
- 599 Bulthuis, D.A., 1987. Effects of temperature on photosynthesis and growth of seagrasses. *Aquat. Bot.* 27, 27–40.  
600 doi:10.1016/0304-3770(87)90084-2
- 601 Carbognin, L., Teatini, P., Tomasin, A., Tosi, L., 2010. Global change and relative sea level rise at Venice: what impact in  
602 term of flooding. *Clim. Dyn.* 35, 1055–1063. doi:DOI 10.1007/s00382-009-0617-5
- 603 Comune di Venezia, 2018. The altimetry of the historical center: percentage of flooding [WWW Document]. URL  
604 <https://www.comune.venezia.it/it/content/le-percentuali-allagamento#ingl> (accessed 10.24.18).
- 605 Consorzio Venezia Nuova, 2018. MOSE [WWW Document]. URL <https://www.mosevenezia.eu/?lang=en/> (accessed  
606 6.12.18).
- 607 D’Alpaos, L., 2010. Fatti e misfatti di idraulica lagunare. Istituto Veneto di Scienze, Lettere ed Arti, Venice, Italy.
- 608 Elliott, M., 2014. Integrated marine science and management: Wading through the morass. *Mar. Pollut. Bull.* 86, 1–4.  
609 doi:10.1016/j.marpolbul.2014.07.026
- 610 Elliott, M., 2011. Marine science and management means tackling exogenic unmanaged pressures and endogenic  
611 managed pressures - A numbered guide. *Mar. Pollut. Bull.* 62, 651–655. doi:10.1016/j.marpolbul.2010.11.033
- 612 Esparza, J., Nielsen, M., 1994. Decidability issues for Petri Nets - a survey. *J. Inform. Process. Cybernet. EIK* 30, 143–  
613 160.
- 614 Fischer, A., Eastwood, A., 2016. Coproduction of ecosystem services as human-nature interactions-An analytical  
615 framework. *Land use policy* 52, 41–50. doi:10.1016/j.landusepol.2015.12.004
- 616 Fongwa, E., Petschick, M., Gnauck, A., Müller, F., 2010. Decision support system for balancing ecosystem services at  
617 the landscape scale: Petri nets modelling application. *Landsc. Ecol.* XXVIII, 241–252.
- 618 Girault, C., Valk, R., 2003. Petri Nets for Systems Engineering. Springer, Berlin, Heidelberg. doi:10.1007/978-3-662-  
619 05324-9
- 620 Heiner, M., Donaldson, R., Gilbert, D., 2010. Petri Nets for Systems Biology. *Symb. Syst. Biol. Theory Methods* 61–97.
- 621 Heiner, M., Gilbert, D., Donaldson, R., 2008. Petri Nets for Systems and Synthetic Biology, in: Bernardo, M., Degano, P.,  
622 Zavattaro, G. (Eds.), *Formal Methods for Computational Systems Biology*. Springer Berlin Heidelberg, Berlin,  
623 Heidelberg, pp. 215–264.
- 624 Heiner, M., Herajy, M., Liu, F., Rohr, C., Schwarick, M., 2012. Snoopy – a unifying Petri net tool. *Proc. PETRI NETS 2012*,  
625 Hamburg, Springer, LNCS 7347, 398–407.
- 626 Jackson, B., Pagella, T., Sinclair, F., Orellana, B., Henshaw, A., Reynolds, B., McIntyre, N., Wheeler, H., Eycott, A., 2013.  
627 Polyscape: A GIS mapping framework providing efficient and spatially explicit landscape-scale valuation of

- 628 multiple ecosystem services. *Landsc. Urban Plan.* 112, 74–88. doi:10.1016/j.landurbplan.2012.12.014
- 629 Jensen, K., 1997. Coloured Petri nets. Basic concepts, analysis methods and practical use. Monographs in Theoretical  
630 Computer Science. Springer, Berlin.
- 631 Lautenbach, S., Kugel, C., Lausch, A., Seppelt, R., 2011. Analysis of historic changes in regional ecosystem service  
632 provisioning using land use data. *Ecol. Indic.* 11, 676–687. doi:10.1016/j.ecolind.2010.09.007
- 633 Levin, S., Xepapadeas, T., Crépin, A.-S., Norberg, J., de Zeeuw, A., Folke, C., Hughes, T., Arrow, K., Barrett, S., Daily, G.,  
634 Ehrlich, P., Kautsky, N., Mäler, K.-G., Polasky, S., Troell, M., Vincent, J.R., Walker, B., 2013. Social-ecological  
635 systems as complex adaptive systems: modeling and policy implications. *Environ. Dev. Econ.* 18, 111–132.  
636 doi:10.1017/S1355770X12000460
- 637 Liu, J., Dietz, T., Carpenter, S.R., Alberti, M., Folke, C., Moran, E., Pell, A.N., Deadman, P., Kratz, T., Lubchenco, J.,  
638 Ostrom, E., Ouyang, Z., Provencher, W., Redman, C.L., Schneider, S.H., Taylor, W.W., 2007. Complexity of  
639 Coupled Human and Natural Systems. *Science (80- )*. 317, 1513–1516. doi:10.1126/science.1144004
- 640 Marani, M., D’Alpaos, A., Lanzoni, S., Carniello, L., Rinaldo, A., 2007. Biologically-controlled multiple equilibria of tidal  
641 landforms and the fate of the Venice lagoon. *Geophys. Res. Lett.* 34. doi:10.1029/2007gl030178
- 642 Martinez-Harms, M.J., Bryan, B.A., Balvanera, P., Law, E.A., Rhodes, J.R., Possingham, H.P., Wilson, K.A., 2015. Making  
643 decisions for managing ecosystem services. *Biol. Conserv.* doi:10.1016/j.biocon.2015.01.024
- 644 McGinnis, M.D., Ostrom, E., 2014. Social-ecological system framework: initial changes and continuing challenges. *Ecol.*  
645 *Soc.* 19. doi:10.5751/es-06387-190230
- 646 Mouchet, M.A., Lamarque, P., Martín-López, B., Crouzat, E., Gos, P., Byczek, C., Lavorel, S., 2014. An interdisciplinary  
647 methodological guide for quantifying associations between ecosystem services. *Glob. Environ. Chang.* 28, 298–  
648 308. doi:10.1016/j.gloenvcha.2014.07.012
- 649 Munari, M., Matozzo, V., Marin, M.G., 2011. Combined effects of temperature and salinity on functional responses of  
650 haemocytetes and survival in air of the clam *Ruditapes philippinarum*. *Fish Shellfish Immunol.* 30, 1024–1030.  
651 doi:10.1016/j.fsi.2011.01.025
- 652 Murata, T., 1989. Petri Nets : Properties , Analysis and Applications. *Proc. IEEE* 77, 541–580.
- 653 Ochoa, V., Urbina-Cardona, N., 2017. Tools for spatially modeling ecosystem services: Publication trends, conceptual  
654 reflections and future challenges. *Ecosyst. Serv.* 26, 155–169. doi:10.1016/j.ecoser.2017.06.011
- 655 Ostrom, E., 2009. A General Framework for Analyzing Sustainability of Social-Ecological Systems. *Science (80- )*. 325,  
656 419–422. doi:10.1126/science.1172133
- 657 Pranovi, F., Caccin, A., Franzoi, P., Malavasi, S., Zucchetta, M., Torricelli, P., 2013. Vulnerability of artisanal fisheries to  
658 climate change in the Venice Lagoon. *J. Fish Biol.* 83, 847–864. doi:10.1111/jfb.12124
- 659 Pranovi, F., Da Ponte, F., Raicevich, S., Giovanardi, O., 2004. A multidisciplinary study of the immediate effects of  
660 mechanical clam harvesting in the Venice Lagoon. *ICES J. Mar. Sci.* 61, 43–52. doi:10.1016/j.icesjms.2003.10.003
- 661 Pranovi, F., Libralato, S., Raicevich, S., Granzotto, A., Pastres, R., Giovanardi, O., 2003. Mechanical clam dredging in  
662 Venice lagoon: ecosystem effects evaluated with a trophic mass-balance model. *Mar. Biol.* 143, 393–403.  
663 doi:10.1007/s00227-003-1072-1
- 664 Rau, A.-L., von Wehrden, H., Abson, D.J., 2018. Temporal Dynamics of Ecosystem Services. *Ecol. Econ.* 151, 122–130.  
665 doi:10.1016/j.ecolecon.2018.05.009
- 666 Ravera, O., 2000. The Lagoon of Venice : the result of both natural factors and human influence. *J. Limnol.* 59, 19–30.
- 667 Reyers, B., Biggs, R., Cumming, G.S., Elmqvist, T., Hejnowicz, A.P., Polasky, S., 2013. Getting the measure of ecosystem  
668 services: a social–ecological approach. *Front. Ecol. Environ.* 11, 268–273. doi:10.1890/120144
- 669 Rieb, J.T., Chaplin-Kramer, R., Daily, G.C., Armsworth, P.R., Böhning-Gaese, K., Bonn, A., Cumming, G.S., Eigenbrod, F.,  
670 Grimm, V., Jackson, B.M., Marques, A., Pattanayak, S.K., Pereira, H.M., Peterson, G.D., Ricketts, T.H., Robinson,  
671 B.E., Schröter, M., Schulte, L.A., Seppelt, R., Turner, M.G., Bennett, E.M., 2017. When, Where, and How Nature  
672 Matters for Ecosystem Services: Challenges for the Next Generation of Ecosystem Service Models. *Bioscience* 67,

- 673 820–833. doi:10.1093/biosci/bix075
- 674 Rizzetto, F., Tosi, L., 2011. Aptitude of modern salt marshes to counteract relative sea-level rise, Venice Lagoon (Italy).  
675 *Geology* 39, 755–758. doi:10.1130/g31736.1
- 676 Rova, S., Pranovi, F., 2017. Analysis and management of multiple ecosystem services within a social-ecological context.  
677 *Ecol. Indic.* 72, 436–443. doi:10.1016/j.ecolind.2016.07.050
- 678 Rova, S., Pranovi, F., Müller, F., 2015. Provision of ecosystem services in the lagoon of Venice (Italy): an initial spatial  
679 assessment. *Ecohydrol. Hydrobiol.* 15, 13–25. doi:10.1016/j.ecohyd.2014.12.001
- 680 Rukundo, E., Liu, S., Dong, Y., Rutebuka, E., Asamoah, E.F., Xu, J., Wu, X., 2018. Spatio-temporal dynamics of critical  
681 ecosystem services in response to agricultural expansion in Rwanda, East Africa. *Ecol. Indic.* 89, 696–705.  
682 doi:10.1016/j.ecolind.2018.02.032
- 683 Saunders, M.I., Leon, J., Phinn, S.R., Callaghan, D.P., O’Brien, K.R., Roelfsema, C.M., Lovelock, C.E., Lyons, M.B.,  
684 Mumby, P.J., 2013. Coastal retreat and improved water quality mitigate losses of seagrass from sea level rise.  
685 *Glob. Chang. Biol.* 19, 2569–2583. doi:10.1111/gcb.12218
- 686 Schlüter, M., Hinkel, J., Bots, P.W.G., Arlinghaus, R., 2014. Application of the SES Framework for Model-based Analysis  
687 of the Dynamics of Social-Ecological Systems. *Ecol. Soc.* 19. doi:10.5751/es-05782-190136
- 688 Sharp, R., Tallis, H.T., Ricketts, T., Guerry, A.D., Wood, S.A., Chaplin-Kramer, R., Nelson, E., Ennaanay, D., Wolny, S.,  
689 Olwero, N., Vigerstol, K., Pennington, D., Mendoza, G., Aukema, J., Foster, J., Forrest, J., Cameron, D., Arkema,  
690 K., Lonsdorf, E., Kennedy, C., Verutes, G., Kim, C.K., Guannel, G., Papenfus, M., Toft, J., Marsik, M., Bernhardt, J.,  
691 Griffin, R., Glowinski, K., Chaumont, N., Perelman, A., Lacayo, M., Mandle, L., Griffin, R., Hamel, P., 2014. InVEST  
692 User’s Guide. Nat. Cap. Proj. Stanford.
- 693 Snoopy, 2017. Snoopy website [WWW Document]. URL [http://www-dssz.informatik.tu-](http://www-dssz.informatik.tu-cottbus.de/DSSZ/Software/Snoopy)  
694 [cottbus.de/DSSZ/Software/Snoopy](http://www-dssz.informatik.tu-cottbus.de/DSSZ/Software/Snoopy) (accessed 6.14.18).
- 695 Solidoro, C., Bandelj, V., Aubry Bernardi, F., Camatti, E., Ciavatta, S., Cossarini, G., Facca, C., Franzoi, P., Libralato, S.,  
696 Melaku Canu, D., Pastres, R., Pranovi, F., Raicevich, S., Socal, G., Sfriso, A., Sigovini, M., Tagliapietra, D., Torricelli,  
697 P., 2010. Response of the Venice Lagoon Ecosystem to Natural and Anthropogenic Pressures over the Last 50  
698 Years, in: Kennish, M.J., Paerl, H.W. (Eds.), *Coastal Lagoons: Critical Habitats of Environmental Change*. CRC  
699 press, pp. 483–511.
- 700 Spake, R., Lasseur, R., Crouzat, E., Bullock, J.M., Lavorel, S., Parks, K.E., Schaafsma, M., Bennett, E.M., Maes, J.,  
701 Mulligan, M., Mouchet, M., Peterson, G.D., Schulp, C.J.E., Thuiller, W., Turner, M.G., Verburg, P.H., Eigenbrod, F.,  
702 2017. Unpacking ecosystem service bundles: Towards predictive mapping of synergies and trade-offs between  
703 ecosystem services. *Glob. Environ. Chang.* 47, 37–50. doi:10.1016/j.gloenvcha.2017.08.004
- 704 Stürck, J., Schulp, C.J.E., Verburg, P.H., 2015. Spatio-temporal dynamics of regulating ecosystem services in Europe-  
705 The role of past and future land use change. *Appl. Geogr.* 63, 121–135. doi:10.1016/j.apgeog.2015.06.009
- 706 Tallis, H., Polasky, S., 2009. Mapping and valuing ecosystem services as an approach for conservation and natural-  
707 resource management. *Ann N Y Acad Sci* 1162, 265–283. doi:10.1111/j.1749-6632.2009.04152.x
- 708 Umgiesser, G., Matticchio, B., 2006. Simulating the mobile barrier (MOSE) operation in the Venice Lagoon, Italy: global  
709 sea level rise and its implication for navigation. *Ocean Dyn.* 56, 320–332. doi:DOI 10.1007/s10236-006-0071-4
- 710 Velez, C., Figueira, E., Soares, A.M.V.M., Freitas, R., 2017. Effects of seawater temperature increase on economically  
711 relevant native and introduced clam species. *Mar. Environ. Res.* 123, 62–70.  
712 doi:10.1016/j.marenvres.2016.11.010
- 713 Villa, F., Bagstad, K.J., Voigt, B., Johnson, G.W., Portela, R., Honzak, M., Batker, D., 2014. A Methodology for Adaptable  
714 and Robust Ecosystem Services Assessment. *PLoS One* 9. doi:ARTN e91001 DOI 10.1371/journal.pone.0091001
- 715 Voulvoulis, N., Arpon, K.D., Giakoumis, T., 2017. The EU Water Framework Directive: From great expectations to  
716 problems with implementation. *Sci. Total Environ.* 575, 358–366. doi:10.1016/j.scitotenv.2016.09.228
- 717 Xu, X., Yang, G., Tan, Y., Liu, J., Hu, H., 2018. Ecosystem services trade-offs and determinants in China’s Yangtze River  
718 Economic Belt from 2000 to 2015. *Sci. Total Environ.* 634, 1601–1614. doi:10.1016/j.scitotenv.2018.04.046



720 FIGURES  
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723 **Figure 1.** General Petri net structure developed for modeling ecosystem services (ES) with direct flow type (regulating ES, A) and with  
724 mediated flow type (provisioning and cultural ES, B). Circles = places (i.e. elements of the system); squares = transitions (ecosystem  
725 functions, activities, interactions); solid arrows = normal arcs (i.e. transitions consume the elements in the input places); solid lines  
726 ending with a circle = read arcs (i.e. elements in the input places are needed but not consumed by the transition); dashed arrows =  
727 modifier arcs (i.e. input places can modify the rate of the transition but are not a precondition for the transition).

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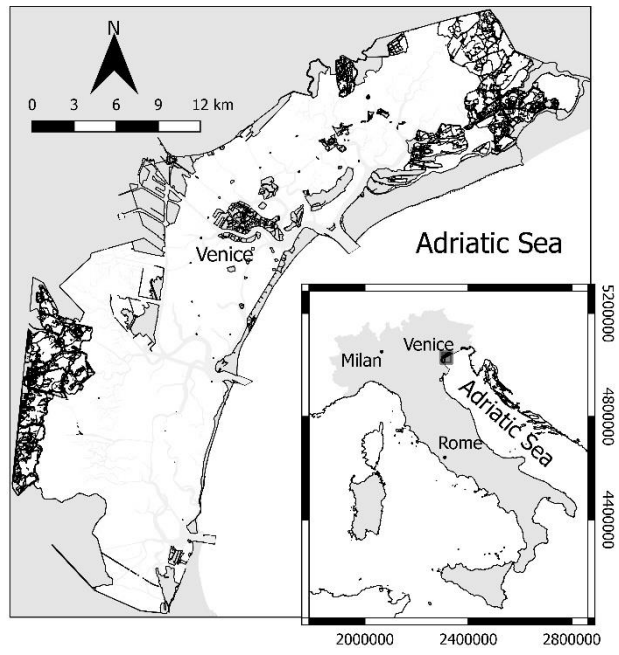


Figure 2. Case study area: the Venice lagoon (Italy).

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ES flow type	Resource system(s)	Negative externalities	ES groups
Direct	Habitats	No	<ul style="list-style-type: none"> <li>- Climate regulation;</li> <li>- Waste treatment;</li> <li>- Erosion prevention 1;</li> <li>- Erosion prevention 2;</li> <li>- Lifecycle maintenance</li> </ul> <b>Regulating</b>
		Yes	
Mediated	Fauna	No	<ul style="list-style-type: none"> <li>- Artisanal fishing;</li> <li>- Recreational fishing;</li> <li>- Hunting</li> </ul> <b>Provisioning 1</b>
		Yes	<ul style="list-style-type: none"> <li>- Clam harvesting</li> </ul> <b>Provisioning 2</b>
	Habitats + Heritage + Channels	No	<ul style="list-style-type: none"> <li>- Info. for cogn. dev.;</li> <li>- Traditions</li> </ul> <b>Cultural 1</b>
		Yes	<ul style="list-style-type: none"> <li>- Tourism</li> </ul> <b>Cultural 2</b>
		Yes	<ul style="list-style-type: none"> <li>- Navigation</li> </ul> <b>Cultural 3</b>
Channels	Yes		

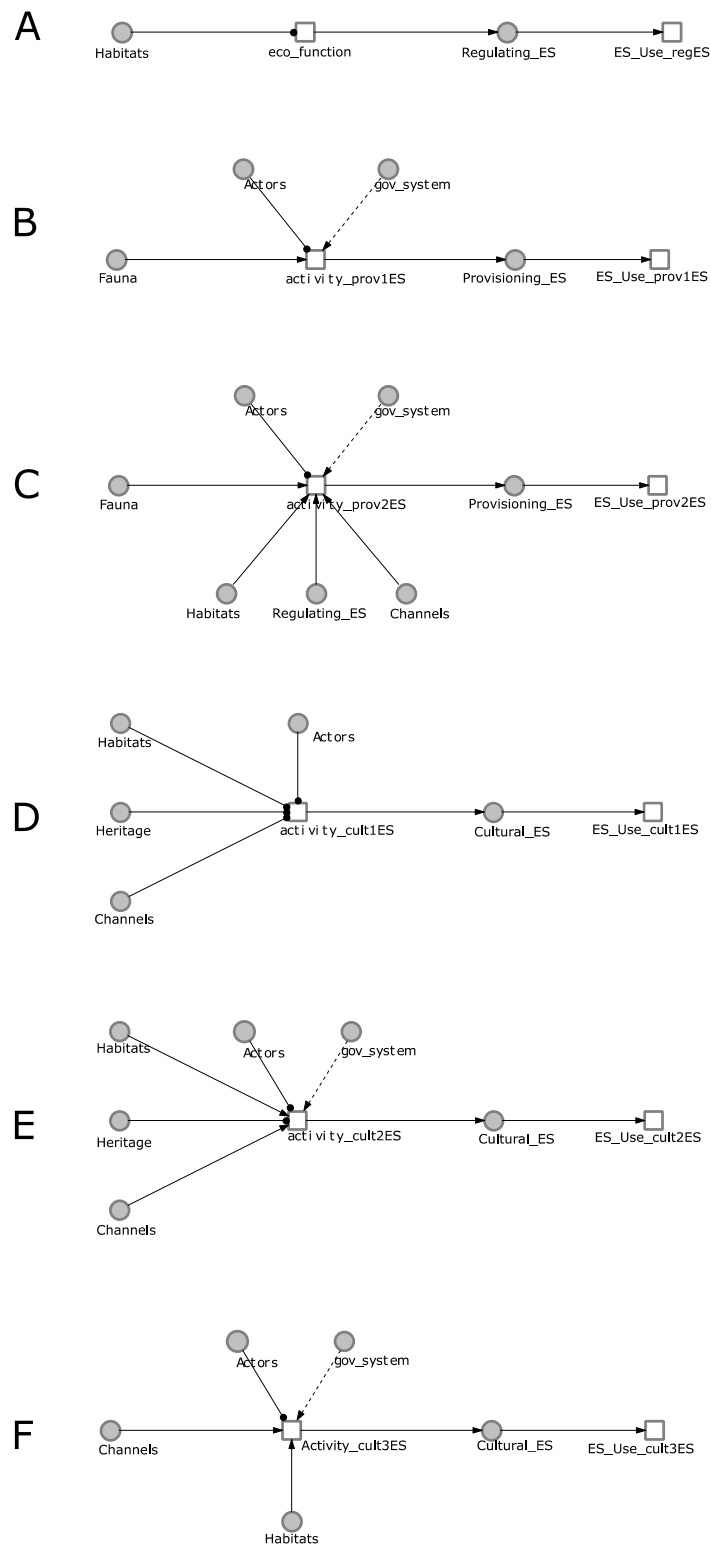
**Figure 3.** Logical flow diagram for the definition of the ecosystem services (ES) “topological” groups

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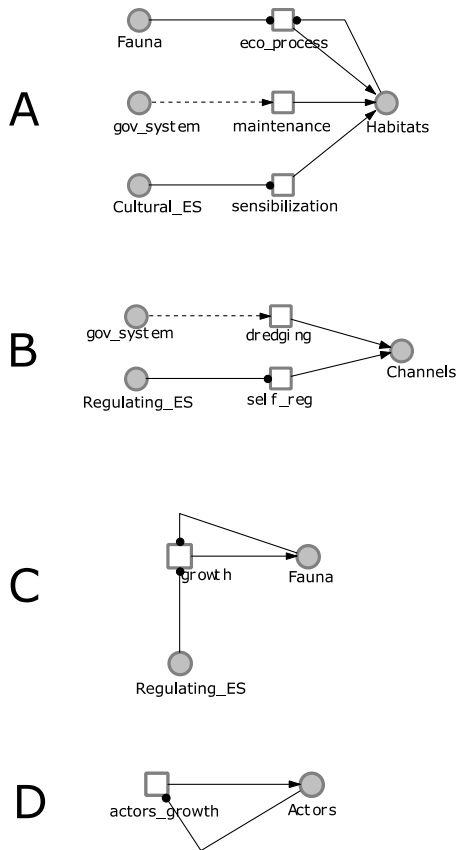
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737 **Figure 4.** Graphical structure of the six ecosystem services (ES) subnets (regulating 1 ES (A), provisioning 1 ES (B), provisioning 2 ES  
 738 (C), cultural 1 ES (D), cultural 2 ES (E), cultural 3 ES (F)). Circles = places; squares = transitions; solid arrows = normal arcs; solid lines  
 739 ending with a circle = read arcs; dashed arrows = modifier arcs.

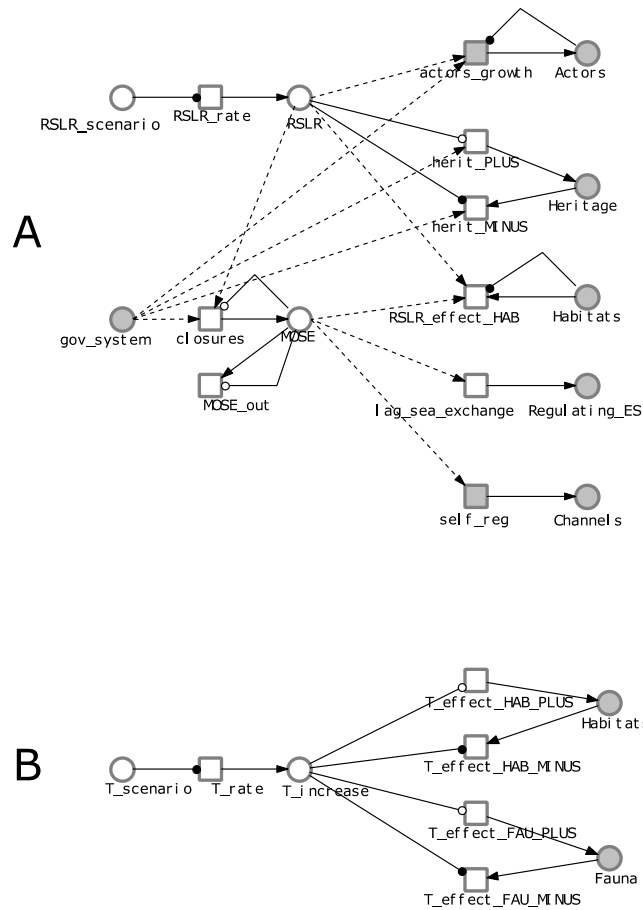




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741 **Figure 5.** Graphical structure of the subnets representing the processes generating the resource units (habitats (A), channels (B) and  
 742 fauna (C)), and actor's growth (D). Circles = places; squares = transitions; solid arrows = normal arcs; solid lines ending with a circle  
 743 = read arcs; dashed arrows = modifier arcs.

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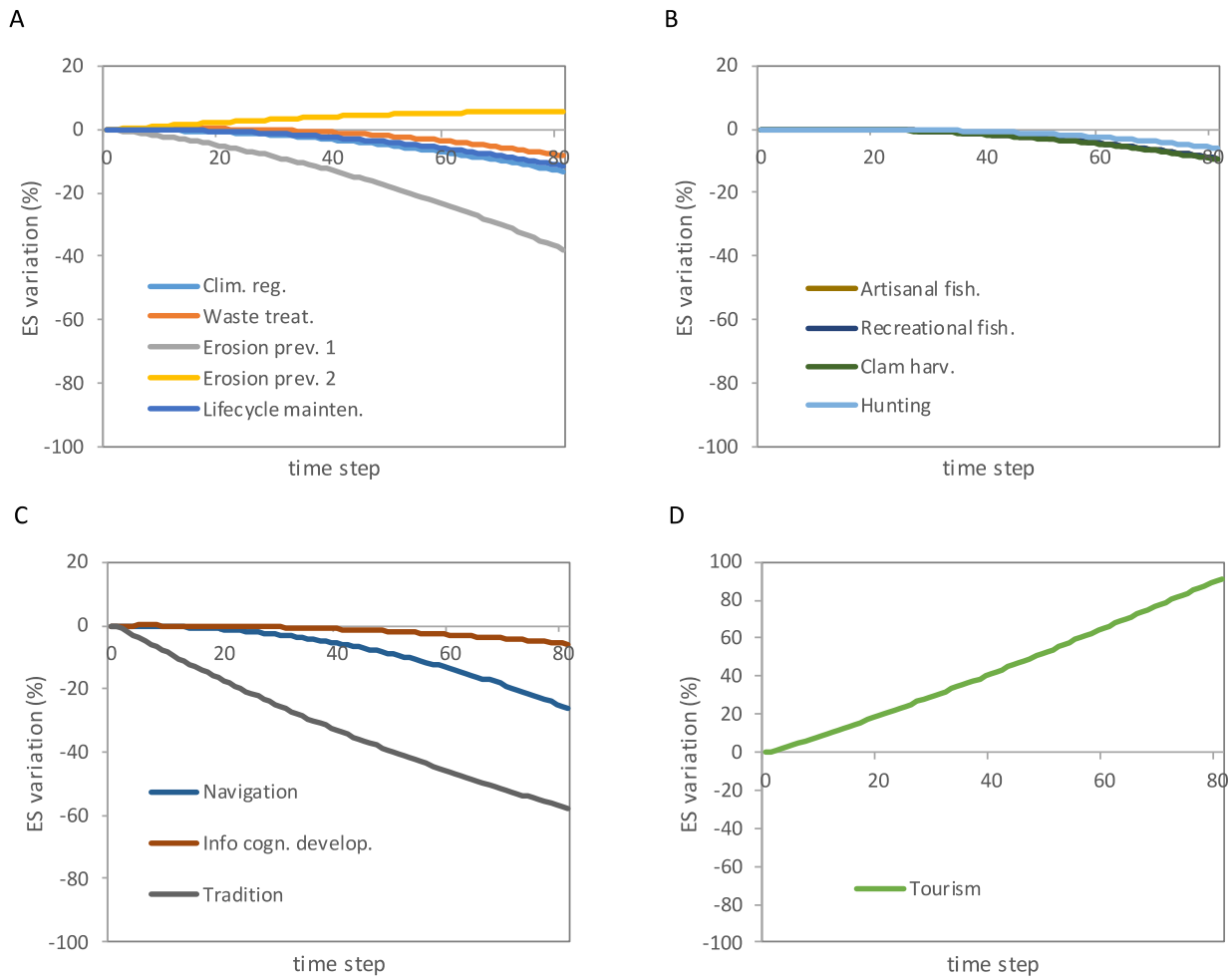
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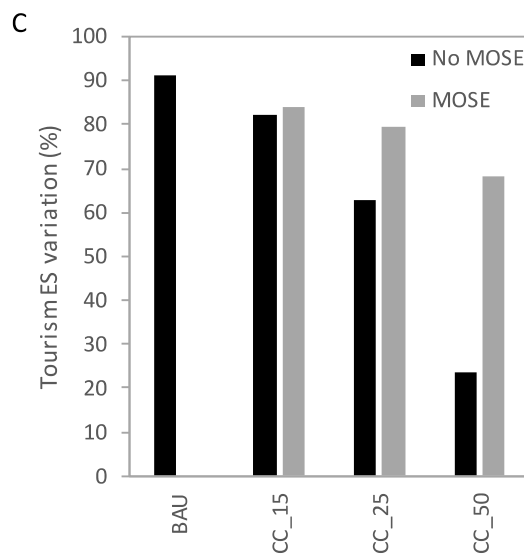
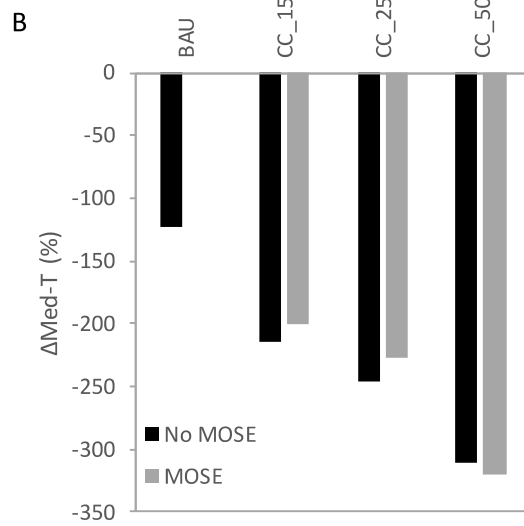
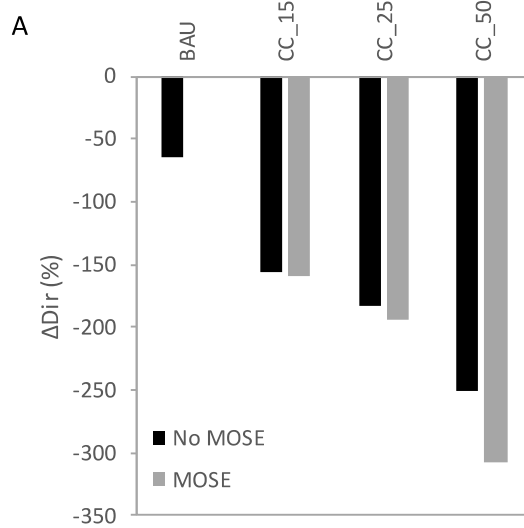
**Figure 6.** Graphical structure of the subnets modeling the effects of relative sea level rise (RSLR) and MOSE system (A) and of temperature (T) increase (B). The following additional model variables were added to model these effects: RSLR\_scenario: specifies the RSLR scenario (none, +15 cm, +25 cm and +50 cm by the end of the 21st century); RSLR: RSLR at each time step; MOSE: n. of MOSE closures per year at each time step; T\_scenario: specifies the T scenario (none, +1°C by the end of the 21st century); T\_increase: T increase at each time step. For a detailed description of how these subnets work please refer to Appendix A. Circles = places; squares = transitions; solid arrows = normal arcs; solid lines ending with a circle = read arcs; dashed arrows = modifier arcs.



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**Figure 7.** Ecosystem services (ES) variation (%) over time under Business-As-Usual (BAU) scenario. Regulating ES (A), provisioning ES (B), cultural ES except tourism (C), tourism (D).

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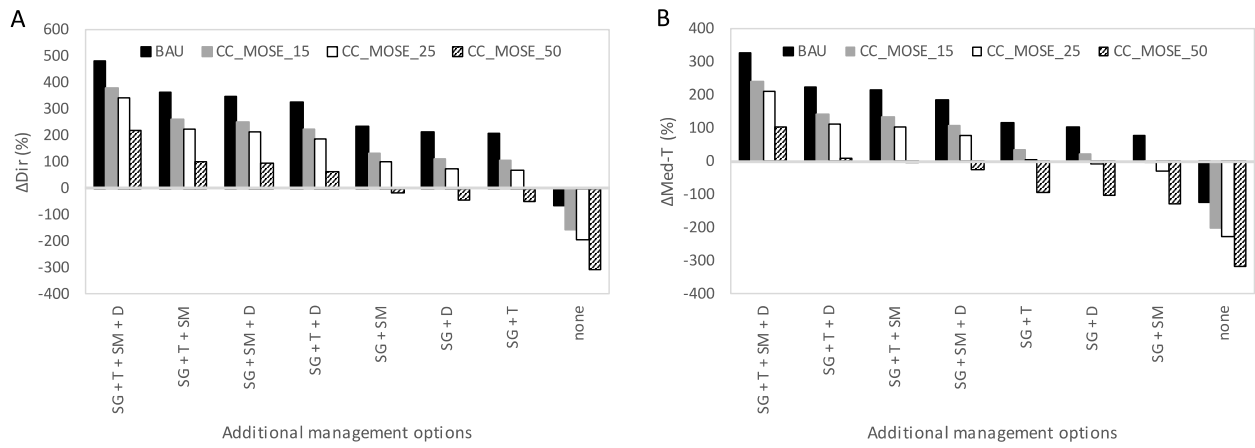
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**Figure 8.** Values assumed by the two aggregated indicators  $\Delta Dir$ (A) and  $\Delta Med-T$  (B), and variation of the Tourism ES (C), at the end of 21st century under the Business-as-Usual (BAU) and climate change (CC) scenarios, with and without functioning of the MOSE system.



760 **Figure 9.** Effect of combined additional management options with respect to the two aggregated indicators  $\Delta Dir$  (A) and  $\Delta Med-T$  (B)  
 761 under the BAU and CC\_MOSE scenarios. Abbreviations: SG, seagrasses maintenance; SM, salt marshes maintenance; D, benthic  
 762 diatoms maintenance; T, tourism.  
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## TABLES

**Table 1.** Ecosystem services (ES) included in the model and their indicators. More details on the modeling of each ES are provided in Appendix A.

ES category	ES	Indicator
Regulating	Climate regulation	Amount of carbon sequestered by seagrasses and salt marshes habitats
	Waste treatment	Self-depuration capacity indicated through the amount of nitrogen removed through denitrification
	Erosion prevention 1	Areas in which salt marshes provide a sheltering effect with respect to wind driven erosion
	Erosion prevention 2	Sum of habitats' biostabilization capacity, that reduces the bottoms' susceptibility to erosion
	Lifecycle maintenance	Sum of habitats' nursery role
Provisioning	Artisanal fishing	Yield from artisanal fishing activities
	Recreational fishing	Yield from recreational fishing activities
	Hunting	Yield from recreational bird hunting activities
	Clam harvesting	Yield mechanical clam harvesting activities
Cultural	Info. for cognitive development	n. of visitors through environmental education activities
	Traditions	n. of people practicing traditional activities
	Tourism	n. of visitors to the lagoon (historical center of Venice excluded)
	Navigation	n. of recreational boats' passages

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**Table 2.** Colorsets (italics) and colors (numbered elements) representing the social-ecological system's (SES) elements involved in the model

<b>Resource systems (colorsets) and resource units (colors within the colorsets)</b>			
<i>Habitats</i>	<i>Fauna</i>	<i>Channels</i>	<i>Resources deriving from past states of the SES (Heritage)</i>
- 0 Salt marshes	- 0 Target fish species	- 0 Channels	- 0 Density of cultural heritage
- 1 Seagrasses	- 1 Clams		- 1 Traditional knowledge
- 2 Bare (intertidal)	- 2 Birds		
- 3 Benthic diatoms			
- 4 Macroalgae			
<b>ES categories (colorsets) and ES (colors within the colorsets)</b>			
<i>Regulating ES</i>	<i>Provisioning ES (*)</i>		<i>Cultural ES</i>
- 0 Climate regulation	- 0 Artisanal fishing		- 0 Tourism
- 1 Waste treatment	- 1 Recreational fishing		- 1 Navigation
- 2 Erosion prevention 1	- 2 Clam harvesting		- 2 Information for cognitive development
- 3 Erosion prevention 2	- 3 Hunting		- 3 Traditions
- 4 Lifecycle maintenance			
<b>Governance system (colorset) and management fields (colors within the colorset)</b>			
<i>Governance system</i>			
- 0 Tourism			
- 1 Navigation			
- 2 Artisanal fishing			
- 3 Recreational fishing			
- 4 Clam harvesting			
- 5 Hunting			
- 6 Salt marsh maintenance			
- 7 Seagrass maintenance			
- 8 Bare (intertidal) maintenance			
- 9 Benthic diatoms maintenance			
- 10 Macroalgae maintenance			
- 11 Channel dredging			
- 12 Lagoon-sea exchanges			
<b>Actors (colorset) and types of actors (colors within the colorset)</b>			
<i>Actors</i>			
- 0 Residents			
- 1 Artisanal fishermen			
- 2 Recreational fishermen			
- 3 Clam fishermen			
- 4 Hunters			
- 5 Users of environmental education activities			
- 6 Tourists			
- 7 Boat owners			

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(\*) Recreational fishing and hunting are here classified as provisioning ES as they yield tangible products, but can be also assimilated to cultural ES due to their recreational importance.

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**Table 3.** Unfolding of the places involved in the generation of each ecosystem service (ES). Abbreviations of the ES “topological” groups: R, regulating; P1, provisioning 1; P2, provisioning 2; C1, cultural 1; C2, cultural 2; C3, cultural 3.

ES group	ES	Habitats resource units	Fauna resource units	Channels resource units	Heritage resource units	Actors	Governance system’s management fields	ES
R (Fig. 2A)	Climate regulation	Salt marshes Seagrasses						
R (Fig. 2A)	Waste treatment	Seagrasses Benthic diatoms Bare (intertidal) Macroalgae						
R (Fig. 2A)	Erosion prevention 1	Salt marshes						
R (Fig. 2A)	Erosion prevention 2	Seagrasses Benthic diatoms Macroalgae						
R (Fig. 2A)	Lifecycle maintenance	ALL						
P1 (Fig. 2B)	Artisanal fishing		Target fish species			Artisanal fishermen	Artisanal fishing	
P1 (Fig. 2B)	Recreational fishing		Target fish species			Recreational fishermen	Recreational fishing	
P1 (Fig. 2B)	Hunting		Birds			Hunters	Hunting	
P2 (Fig. 2C)	Clam harvesting	Seagrasses (*) Benthic diatoms (*)	Clams	Channels (*)		Clam fishermen	Clam harvesting	Lifecycle maintenance (*)
C1 (Fig. 2D)	Info. for cognitive development	ALL		Channels	Density of cultural heritage	Users of environmental education activities		
C1 (Fig. 2D)	Traditions	ALL		Channels	Traditional knowledge	Residents		
C2 (Fig. 2E)	Tourism	Salt marshes(*) Seagrasses(*) Bare (intertidal)(*) Benthic diatoms(*)		Channels (*)	Density of cultural heritage	Tourists	Tourism	
C3 (Fig. 2F)	Navigation	Salt marshes(*) Seagrasses(*) Bare (intertidal)(*) Benthic diatoms(*)		Channels (*)		Boat owners	Navigation	

777 (\*) places impacted through negative externalities



778 **Table 4.** Ranking of management options with respect to the two aggregated indicators  $\Delta\text{Dir}$  and  $\Delta\text{Med-T}$ . The ranking  
 779 is the same in the BAU and CC\_MOSE scenarios, except for the groups of options marked with (\*) and (\*\*) (grey  
 780 background), for which the relative ranking varies between scenarios.

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$\Delta\text{Dir}$	$\Delta\text{Med-T}$
1 seagrass mainten.	1 seagrass mainten.
2 salt marsh mainten.	2 tourism
3 benthic diatoms mainten.	3 benthic diatoms mainten.
4 tourism	4 bare (intertidal) mainten.
5 navigation (*)	5 artisanal fishing(**)
5 bare (intertidal) mainten. (*)	5 salt marsh mainten. (**)
7 hunting	5 hunting (**)
8 artisanal fishing	5 macroalgae mainten. (**)
9 macroalgae mainten.	5 navigation (**)
10 recreational fishing	10 channels' dredging
11 clam harvesting	11 recreational fishing
12 channels' dredging	12 clam harvesting

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**Table 5.** Model results under the BAU and CC\_MOSE scenarios (expressed as ecosystem services (ES) variation (%) at the end of the 21<sup>st</sup> century), obtained with the complete model and with the three additional configurations created to test the effects of neglecting the interactions among ES.

Scenario	Model configuration	Clim. reg.	Waste treat.	Erosion prev. 1	Erosion prev. 2	Lifecycle maint.	Tourism	Navigation	Info. cogn. dev.	Tradition	Artisanal fish.	Recreat. fish.	Clam harv.	Hunting	ΔDir	ΔMed-T
BAU	complete	-13	-8	-38	6	-11	91	-26	-6	-58	-9	-9	-10	-6	-65	-124
	NO_EcoFeedbacks	-13	-8	-37	6	-12	91	-27	-6	-58	0	0	0	0	-64	-91
	NO_ES_sideEffects	5	6	-19	18	4	110	0	2	-54	3	3	3	2	15	-41
	NO_ALL	5	6	-19	18	3	110	0	2	-54	0	0	0	0	13	-52
CC_MOSE_15	complete	-41	-23	-47	-21	-27	84	-32	-14	-61	-27	-27	-27	-14	-158	-201
	NO_EcoFeedbacks	-40	-22	-46	-20	-28	84	-35	-14	-61	-5	-5	-4	0	-157	-124
	NO_ES_sideEffects	-25	-10	-29	-10	-15	103	-2	-6	-58	-17	-17	-17	-7	-89	-123
	NO_ALL	-24	-9	-29	-9	-15	103	-2	-6	-58	-5	-5	-4	0	-85	-79
CC_MOSE_25	complete	-51	-29	-53	-28	-34	80	-35	-17	-63	-32	-32	-32	-18	-194	-228
	NO_EcoFeedbacks	-50	-28	-52	-27	-35	80	-38	-17	-63	-5	-5	-4	0	-192	-131
	NO_ES_sideEffects	-36	-16	-36	-18	-23	98	-3	-9	-59	-23	-23	-23	-11	-129	-152
	NO_ALL	-34	-15	-35	-17	-22	98	-3	-9	-59	-5	-5	-4	0	-124	-85
CC_MOSE_50	complete	-78	-51	-67	-53	-58	68	-42	-26	-67	-51	-51	-52	-32	-307	-320
	NO_EcoFeedbacks	-77	-50	-66	-53	-62	68	-48	-26	-67	-5	-5	-4	0	-307	-155
	NO_ES_sideEffects	-65	-41	-52	-45	-52	85	-12	-19	-64	-45	-45	-46	-27	-256	-257
	NO_ALL	-64	-39	-52	-43	-50	85	-12	-19	-64	-5	-5	-4	0	-247	-109

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