

Manuscript Number:

Title: Photo-oxidation products of  $\alpha$ -pinene in coarse, fine and ultrafine aerosol: a new high sensitive HPLC-MS/MS method

Article Type: Research Paper

Keywords: aerosol; ultrafine particles; HPLC-MS/MS; cis-pinonic acid; pinic acid.

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Abstract: Oxidation products of  $\alpha$ -pinene represent a fraction of organic matter in the environmental aerosol.  $\alpha$ -pinene is one of most abundant monoterpenes released in the atmosphere by plants, located typically in boreal, temperate and tropical forests. This primary compound reacts with atmospheric oxidants, such as O<sub>3</sub>, O<sub>2</sub>, OH radicals and NO<sub>x</sub>, through the major tropospheric degradation pathway for many monoterpenes under typical atmospheric condition. Although several studies identified a series of by-products deriving from the  $\alpha$ -pinene photo-oxidation in the atmosphere, such as pinic and cis-pinonic acid, the knowledge of the mechanism of this process is partially still lacking. Thus, the investigation of the distribution of these acids in the different size aerosol particles provide additional information on this regard. The aim of this study is twofold. First, we aim to improve the existing analytical methods for the determination of pinic and cis-pinonic acid in aerosol samples, especially in terms of analytical sensitivity and limits of detection (LOD) and quantification (LOQ). We even attempted to increase the knowledge of the  $\alpha$ -pinene photo-oxidation processes by analysing, for the first time, the particle-size distribution up to nanoparticle level of pinic and cis-pinonic acid. The analysis of aerosol samples was carried out via high-performance liquid chromatography coupled to a triple quadrupole mass spectrometer. The instrumental LOD values of cis-pinonic and pinic acid are 1.6 and 1.2 ng L<sup>-1</sup> while LOQ values are 5.4 and 4.1 ng L<sup>-1</sup>, respectively. Samples were collected by MOUDI IITM cascade impactor with twelve cut-sizes, from March to May 2016 in the urban area of Mestre-Venice (Italy). The range concentrations in the aerosol samples were from 0.6 to 0.8 ng m<sup>-3</sup> for cis-pinonic acid and from 0.1 to 0.8 ng m<sup>-3</sup> for pinic acid.

**Photo-oxidation products of  $\alpha$ -pinene in coarse, fine and ultrafine aerosol: a new high sensitive HPLC-MS/MS method**

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Dear Editor,

We have the pleasure to send you by electronic submission a copy of the work titled "Photo-oxidation products of  $\alpha$ -pinene in coarse, fine and ultrafine aerosol: a new high sensitive HPLC-MS/MS method" by Matteo Feltracco, Elena Barbaro, Daniele Contini, Roberta Zangrando, Giuseppa Toscano, Dario Battistel, Carlo Barbante, Andrea Gambaro to be considered for publication in Atmospheric Environment.

Best regards,

Matteo Feltracco

The purpose of this work is to present a high sensitive method to determine cis-pinonic and pinic acids in different particle size fractions including, for the first time, the nanoparticles. The quantitative performance of HPLC coupled to triple quadrupole API 4000 were carried out to

determine these acids in environmental samples at trace concentration levels. The article presents a theoretical section regarding analysis and there is a wide contribute to each phase of analytical operations. An important improvement of this work is to develop a higher sensitive analytical method than previous studies, suitable for an evaluation of the particle-size distribution of the target compounds in aerosols.

Supporting material (SM) include material directly relevant to the conclusion of a paper that cannot be included. SM provides figures and tables that enrich the data described in the paper.

This article has not yet been published and it is not under consideration by any other journal. All authors are aware of the manuscript and they accept responsibility for it.

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3  
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21

22 **Abstract**

23 Oxidation products of  $\alpha$ -pinene represent a fraction of organic matter in the environmental aerosol.  $\alpha$ -  
24 pinene is one of most abundant monoterpenes released in the atmosphere by plants, located typically in  
25 boreal, temperate and tropical forests. This primary compound reacts with atmospheric oxidants, such  
26 as  $O_3$ ,  $O_2$ , OH radicals and  $NO_x$ , through the major tropospheric degradation pathway for many  
27 monoterpenes under typical atmospheric condition. Although several studies identified a series of by-  
28 products deriving from the  $\alpha$ -pinene photo-oxidation in the atmosphere, such as pinic and cis-pinonic  
29 acid, the knowledge of the mechanism of this process is partially still lacking. Thus, the investigation of  
30 the distribution of these acids in the different size aerosol particles provides additional information on  
31 this regard.

32 The aim of this study is twofold. First, we aim to improve the existing analytical methods for the  
33 determination of pinic and cis-pinonic acid in aerosol samples, especially in terms of analytical  
34 sensitivity and limits of detection (LOD) and quantification (LOQ). We even attempted to increase the  
35 knowledge of the  $\alpha$ -pinene photo-oxidation processes by analysing, for the first time, the particle-size  
36 distribution up to nanoparticle level of pinic and cis-pinonic acid. The analysis of aerosol samples was  
37 carried out via high-performance liquid chromatography coupled to a triple quadrupole mass  
38 spectrometer. The instrumental LOD values of cis-pinonic and pinic acid are 1.6 and 1.2  $ng\ L^{-1}$  while  
39 LOQ values are 5.4 and 4.1  $ng\ L^{-1}$ , respectively. Samples were collected by MOUDI II<sup>TM</sup> cascade  
40 impactor with twelve cut-sizes, from March to May 2016 in the urban area of Mestre-Venice (Italy).

41 The range concentrations in the aerosol samples were from 0.6 to 0.8  $ng\ m^{-3}$  for cis-pinonic acid and  
42 from 0.1 to 0.8  $ng\ m^{-3}$  for pinic acid.

## 43 **Introduction**

44 The formation of secondary organic aerosols (SOA) in the rural atmosphere has attracted growing  
45 interest in recent years. The atmospheric formation of new particles and their chemical composition can  
46 have an important role for the determination of the global aerosol load and their effect on the climate  
47 change (Kulmala et al., 2004).

48 Monoterpenes are the most abundant biogenic hydrocarbons in troposphere and these compounds affect  
49 the oxidising capacity of the atmosphere (Kanakidou et al., 2000). In the last years, considerable studies  
50 (Lamb et al., 1987; Larsen et al., 1999; Librando and Tringali, 2005) have been carried out to determine  
51 the secondary organic aerosol (SOA) formation from the photo-oxidation of volatile organic  
52 compounds (VOCs). In particular,  $\alpha$ -pinene is the most important monoterpene released by biogenic  
53 sources, particularly conifers. It has an emission rate projected at global scale of about 127 Tg y<sup>-1</sup>  
54 (Guenther et al., 1995) and it has been shown to give high SOA rates in laboratory smog chamber  
55 research (Lamb et al., 1987; Larsen et al., 1999). The emission on global scale of total monoterpenes  
56 from the vegetation has been estimated between 120–480 Tg y<sup>-1</sup> (Fehsenfeld et al., 1992), therefore the  
57  $\alpha$ -pinene fraction represent a large part of global monoterpenes, by considering its emission rate.  $\alpha$ -  
58 pinene is a primary ingredient of pine resin and it is also found other conifers and non-coniferous  
59 plants. Nevertheless, the atmospheric degradation proceeds through a very complex mechanism that is  
60 still not totally identified, and this leads to form an abundance of reaction products (Glasius et al., 2000,  
61 1999, Iinuma et al., 2016, 2004; Larsen et al., 1999; Librando and Tringali, 2005; Zhang et al., 1992).

62 The innovative purpose of this work is to present an evaluation of the distribution of the target  
63 compounds in aerosols as a function of particle size focusing on the ultrafine fractions. To our  
64 knowledge, the photo-oxidation products of  $\alpha$ -pinene in the ultrafine fraction have never been  
65 investigated. The quantitative performance of HPLC coupled to triple quadrupole API 4000 were  
66 carried out to determine these acids in environmental samples at trace levels. Recently, photo-oxidation  
67 products of  $\alpha$ -pinene have been investigated using GC-MS and HPLC-MS methods (Ding et al., 2008;  
68 Iinuma et al., 2007; Ion et al., 2005; Parshintsev et al., 2010; Pio et al., 2006; Reinnig et al., 2008;  
69 Sheesley and Kenski, 2004; Zhang et al., 2010). GC-MS is a widely used method for the separation,  
70 identification and quantification of individual organic compounds in aerosol samples, even though low-

71 volatile polar substances have to be derivatized prior to injection (Ion et al., 2005; Szmigielski et al.,  
72 2007) and some compounds might decompose during analysis. HPLC/MS methods were used by some  
73 authors to measure oxidation products from terpenes in atmospheric samples and with chamber  
74 experiments (Anttila et al., 2005; Reinnig et al., 2008; Zhang et al., 2010). Anttila et al. (2005)  
75 investigated the environmental aerosol matter with reversed phase chromatography applied to HPLC  
76 system, coupled to an ion-trap mass spectrometer using an electro-spray ionisation (ESI) interface in  
77 negative ionisation mode. Besides, Renning et al. (2008) carried out the ozonolysis of  $\alpha$ -pinene in a  
78 smog chamber and the samples were investigated using reversed phase chromatography coupled with  
79 ion-trap mass spectrometer though an APCI source operating in positive mode. Zhang et al. (2010),  
80 analysed aerosol samples using C18 column placed in a HPLC system coupled with a hybrid Qq-TOF  
81 mass spectrometer. The benefit of HPLC is the suitability for polar and non-volatile compounds as well  
82 as the conditions through the analysis, while mass spectrometer allows high sensitivity and selectivity.  
83 The main objective of this work is to develop and quantify some of the main photo-oxidation products  
84 of  $\alpha$ -pinene, pinic and pinonic acids, in different particle size fractions of aerosol collected in the urban  
85 area of Mestre-Venice (Italy). A sensitive improvement of analytical method is necessary due to the  
86 high fractionation of aerosol in twelve different size ranges.

## 87 **Experimental section**

### 88 **Reagents and standard solutions**

89 Ultra-grade methanol (MeOH) and ultra-grade acetonitrile (ACN) were purchased from Romil® LTD  
90 (Cambridge, UK), Ultrapure water (18.2 M $\Omega$ , 1 ppb TOC) was produced using a Purelab Ultra System  
91 (Elga®, HighWycombe, UK), formic acid ( $\geq 98\%$ ) eluent additive for HPLC system was obtained from  
92 Fluka (Sigma Aldrich®, Buchs, Switzerland). Cis-pinonic acid (Sigma-Aldrich, Sant Louis, Missouri,  
93 USA) and pinic acid (Santa Cruz Biotechnology®, Dallas, Texas, USA), were prepared by solid  
94 standard (purity  $\geq 98\%$ ) and diluted in ultrapure water. Isotopically labelled vanillin<sup>13</sup>C<sub>6</sub>(VAH\*)  
95 was obtained from Sigma Aldrich®.

96 Ultrapure water produced using a Purelab was furtherly purified to obtain higher water quality  
97 necessary to reach lower limits of detection. The LC-Pak® (Merck KGaA, Darmstadt, Germany)  
98 cartridge can be connected to the outlet of water system to produce water with a TOC level below 1 ppb

99 and deliver ultrapure water with minimum trace organic contamination. The LC-Pak® cartridge uses  
100 the well-known reversed-phase silica purification media to remove traces of neutral organics.

#### 101 **HPLC-ESI-MS/MS**

102 An Agilent 1100 Series HPLC Systems (Waldbronn, Germany) with a binary pump, vacuum degasser,  
103 autosampler and thermostated column compartment) was coupled with an API 4000 Triple Quadrupole  
104 Mass Spectrometer (Applied Biosystem/MSD SCIEX, Concord, Ontario, Canada) using a Turbo V  
105 electrospray source (ESI) that operated in negative mode. The chromatographic separation used for the  
106 sample determination of cis-pinonic acid and pinic acid was conducted using a Zorbax Extend-C18  
107 column (Rapid Resolution, 4.6 -150 mm, 3.5 mm; Agilent Technologies). Elution was achieved by a  
108 linear gradient using as mobile phase water with 0.01% of formic acid (eluent A) and MeOH/ACN  
109 80:20 (eluent B). The binary elution program with flow rate of 0.5 mL min<sup>-1</sup> was as follows: 0-1 min,  
110 20% eluent B; 8-25 min, 100% eluent B; 25-35 min, equilibration with 20% eluent B; 100 µL of sample  
111 was injected for analysis. The mass spectrometer's parameters were set as follows: temperature 650 °C,  
112 ion spray voltage -4500 V, GS1 40 psi, GS2 60 psi, CUR 15 psi, CAD 8 psi and EP -8 V. Data were  
113 collected with multiple reaction monitoring (MRM) mode. The first quadrupole (Q1) selected the  
114 molecular ion, while the third quadrupole (Q3) selected the fragment. Both Q1 and Q3 were set at unit  
115 resolution with peak width of 0.7±0.1 amu at 50% of maximum peak height. To improve the sensitivity,  
116 declustering potential (DP), cell energy (CE) and cell exit potential (CXP) were set, using direct  
117 infusions of 1 mg L<sup>-1</sup> of each individual standard. The voltage of the orifice was controlled by the DP  
118 parameter, the CE was the amount of energy that the precursor ions received as they were accelerated  
119 into the collision cell, and the CXP was used to focus and accelerate the ions after leaving the collision  
120 cell. The monitored transition and instrumental parameters for each compound are shown in Table S-1.  
121 Analyst Software version 1.5.2 (Applied Biosystems MDS SCIEX Instruments) was used for the  
122 identification and quantification of the target compounds.



### 123 **Sample collection**

124 The aerosol samples were collected at Scientific Campus of Ca' Foscari University (45°28'47"N,  
125 12°15'12"E, Mestre-Venice, Italy) during spring 2016, using the rotating model 120 MOUDI-II™  
126 cascade impactor. The particle size distribution was obtained for aerodynamic diameter (D) ranging  
127 from greater of 18 µm to below 0.056 µm. The configuration of sampler consists to eleven stages with  
128 cut sizes at 18, 10, 5.6, 3.2, 1.8, 1.0, 0.56, 0.32, 0.18, 0.10 and 0.056 µm, plus a final back-up filter used  
129 to collect particles with D<0.056 µm. The impaction plates, 25 nm thick aluminium substrate having a  
130 diameter of 47 mm, were prepared in the laboratory and inserted into the impactor at the time of use.  
131 The substrates were made using aluminium foils cut with a hollow cutter. Finally, the collection of  
132 ultrafine atmospheric particles in the last stage was carried out with a quartz fibre filters (SKC Inc.,  
133 Eighty-Four, To-13 model).

134 The duration of the sampling was about 160 h for most of the samples, with an average flow rate of  
135  $30 \pm 1 \text{ L min}^{-1}$ . The air flow at the inlet was measured with a flowmeter before and after each sampling.  
136 The sampling period was chosen to obtain enough matter even for weighting. In fact, one of the most  
137 important feature of MOUDI II™ is the possibility of the weighing of each aluminium substrate, thanks  
138 to their restrained diameter. After conclusion of sampling, the substrates were stored separately at -20  
139 °C until chemical analysis. The sampler was placed on the roof of a building at a height of 20 meters, to  
140 avoid direct human contaminations and to prevent winds modification or alteration.

141 The weather information was provided by station FISTEC-Mestre (IUAV, Venice - Environmental  
142 engineering physics laboratory, website: [fistec.iuav.it](http://fistec.iuav.it)). Precipitation has a great variability especially in  
143 spring due to the high humidity, in fact there were a significant precipitation events from 9<sup>th</sup> to 12<sup>th</sup> of  
144 March (22.8 mm) and a minor from 13<sup>th</sup> to 15<sup>th</sup> of April (4.2 mm) and from 11<sup>th</sup> to 13<sup>th</sup> of May (3.2  
145 mm). The range humidity was included between 22% (4<sup>th</sup> May) and 98.5% (most of the data).  
146 Moreover, the sampling period was conditioned by the medium-high temperature ranging (from 4 °C to  
147 25 °C) and the prevailing winds were from NW and SW with wind speed between 1 and 7 m s<sup>-1</sup> (Figure  
148 S-1 and S-2, supporting material).

149 The impactor with the rotation of the nozzle and impaction plates at 1 rpm, formed a near-uniform  
150 particle deposit on each substrate.

151 **Sample treatment**

152 To determine and quantify cis-pinonic acid and pinic acid, airborne aerosol was collected on aluminium  
153 plates pre-cleaned with MeOH, and on quartz fiber filter, decontaminated with a pre-combustion (4 h at  
154 400 °C in a muffle furnace). Before the closing inside in a clean aluminium foil, the aluminium plates  
155 were weighted to allow the calculation of the collected aerosol.

156 Aluminium plates and filters were removed from the storage package in a laminar flow hood, broken up  
157 into small pieces and placed in a 15 mL vial (previously cleaned with ultra-pure water by sonication at  
158 25 °C) with steel tweezers. 50 µL of isotopically labelled vanillin<sup>13</sup>C<sub>6</sub> (78 ng absolute weight) and 4.95  
159 mL of ultrapure water was added to the substrate before cold-ultrasonically extracting at 10 °C to avoid  
160 the volatilization of the analytes. The extract has been filtrated through a 0.45 mm PTFE filter  
161 (Minisart® Sartorius SRP25, Goettingen, Germany) to remove particulate and filter traces before  
162 instrumental analysis.

163 During the sampling periods, field blanks were taken at the beginning, during and end of the sampling  
164 period. The blank samples were collected by loading, carrying and installing the filter holder in the  
165 instrument with the air pump turned off.

166 **Result and discussion**

167 **Chromatographic separation**

168 The LC–MS/MS conditions were set in terms of eluent composition, eluent pH, gradient conditions and  
169 ionisation conditions using standards pinic acid and cis-pinonic acid. Previous studies demonstrated an  
170 efficient separation of target acids using conventional reversed phase HPLC columns (Anttila et al.,  
171 2005; Parshintsev et al., 2010; Reinnig et al., 2008; Zhang et al., 2010) and GC capillary columns (Ding  
172 et al., 2008; Ion et al., 2005; Pio et al., 2006; Sheesley and Kenski, 2004). The coupling with mass  
173 spectrometer was always preferred to guarantee sensitivity and selectivity.

174 In the present study Zorbax Extend-C18 column 4.6x150 mm (Agilent Technologies, Santa Clara,  
175 California, USA) was chosen for the chromatophic separation and its performance was evaluated by  
176 considering retention time ( $t_R$ ), peak width ( $W$ ), number ( $N$ ) and height (HEPT) of teoretical plates,  
177 asymmetry ( $A_S$ ) for each compound. Resolution ( $R_S$ ) and selectivity ( $\alpha$ ) were evaluated between cis-  
178 pinonic and pinic acid in order to avoid interferences in the transition of each compound.

179 Three different eluent B composition were investigated to improve the chromatographic performance:  
180 pure MeOH (method 1), 80:20 MeOH/ACN (method 2) and 50:50 MeOH/ACN (method 3). MeOH is  
181 the best solvent for the ESI source but the addition of ACN in eluent B can improve the peak symmetry  
182 (Kostiainen and Kauppila, 2009). The composition of eluent A was always ultrapure water with 0.01%  
183 formic acid to improve the peak resolution.

184 Figure 1 shows the chromatographic separations of cis-pinonic and pinic acid and  $^{13}\text{C}_6$  vanillin (internal  
185 standard) using the three different elution program (method 1, 2 and 3).

186 Higher percentage of ACN allowed a more rapid elution of both analytes with method 3 (50:50  
187 ACN:MEOH) respect to pure MeOH used as eluent B (method 1), due to high eluent power of ACN.

188 Moreover, MeOH used as eluent B (method 1) carried out to higher value of peak width for cis-pinonic  
189 acid, while this value decreased by adding ACN in the eluent B (method 2 and 3) (Table S-2). Peak

190 broadening is typically described as plate number (N) or as height equivalent to a theoretical plate  
191 (HETP). This notion is equivalent to a plates-series model reflecting the number of equilibrium steps

192 represented by the column. Table S-1 shows that the method 2 has the higher value of N (and the  
193 smallest HETP value) for cis-pinonic acid, while the method 1 has the higher value of N for pinic acid,

194 although the N value of method 2 is relatively close. Furthermore, another parameter that points out the  
195 good chromatographic conditions is the peak asymmetry. The most important reasons for the presence

196 of asymmetry are slow mass transfer, column overload, the heterogeneity of the stationary phase surface  
197 and the heterogeneity of the column packing (Pápai and Pap, 2002). The method 2 shows an

198 asymmetry closer to 1, approximating in a better way the gaussian form ( $A_S > 1$ : tailing;  $< 1$ : fronting).  
199 Probably, this means that the composition of the mobile phases of the method 2 provided a better mass

200 transfer. We investigate moreover the selectivity ( $\alpha$ ) and the resolution factor ( $R_F$ ). High  $\alpha$  values  
201 indicate good separating power and a good ability of the chromatographic system to distinguish

202 between sample components, and the best values are provided by the method 2 ( $\alpha = 1.12$ ) and 3 ( $\alpha =$   
203 1.20), while the method 1 gave a  $\alpha = 1.07$ . The resolution values are always plenty above 1.5. This

204 ensure that the samples are well separated to a degree at which the area can be accurately measured.  
205 The study of the whole parameters permitted to choose the operative conditions shown in the “HPLC-

206 ESI-MS/MS” section.

207 The chromatographic method is different if compared to some previously published LC–ESI–MS/MS  
208 methods (Anttila et al., 2005; Glasius et al., 1999) in which was used acidic conditions using acetic acid  
209 as buffer and a gradient starting with fully aqueous conditions with methanolic gradient.

210

### 211 **Quantitative performance**

212 The analytical procedure was validated through linear range, instrumental limit of detection (LOD),  
213 instrumental limit of quantification (LOQ), procedural blank, method detection limit (MDL), method  
214 quantification limit (MQL), trueness, repeatability, and extraction yield. The internal standard method  
215 by isotope dilution was used to quantify cis-pinonic and pinic acids and labelled  $^{13}\text{C}_6$  vanillin was  
216 chosen as internal standard because it demonstrated similar instrumental and pre-analytical behaviour.

217 The linearity of the calibration curves for of cis-pinonic acid and pinic acid with labelled  $^{13}\text{C}_6$  vanillin  
218 as internal standard was evaluated using a series of standard solutions prepared in ultrapure water at  
219 average concentrations from 0.01 to 50  $\mu\text{g L}^{-1}$  and a constant concentration of labelled  $^{13}\text{C}_6$  vanillin  
220 ( $15.5 \mu\text{g L}^{-1}$ ). By considering the ratio between concentration of target acids and internal standard and  
221 the ratio between the relative peak areas, linearity was evaluated obtaining  $R^2 \geq 0.9997$ . LOD and LOQ  
222 values are calculated as three and ten times the signal-to-noise ratio of the known absolute amounts of  
223 the analysed target compound in a standard solution (Bliesner, 2006). The LOD values of cis-pinonic  
224 and pinic acid are 1.6 and 1.2  $\text{ng L}^{-1}$  while LOQ values are 5.4 and 4.1  $\text{ng L}^{-1}$ , respectively. Parshintsev  
225 et al. (Parshintsev et al., 2010) obtained a LOD value of 27 and 12  $\mu\text{g L}^{-1}$ , while Zhang et al. (Zhang et  
226 al., 2010) obtained 3.3 and 0.35  $\mu\text{g L}^{-1}$ , respectively for cis-pinonic acid and pinic acid. In both of cases  
227 the studies were carried out via HPLC-MS systems and the LOD values of this study are considerable  
228 lower than reported literature. Furthermore, to our knowledge, the method has the lowest LOD values  
229 compared to previous studies. The instrumental precision was evaluated and CV% value (reported as a  
230 percentage and calculated from the average and standard deviation, calculated as  $(\text{SD}/\text{A}) \times 100$ ) was  
231 below of 10%. Due to the lack of certified reference materials for cis-pinonic and pinic acid in the  
232 aerosol or dust, we estimated trueness, precision and recovery by analyzing five spiked cleaned  
233 aluminium plate and QFF with 63 ng of cis-pinonic acid, 45 ng of pinic acid and 78 ng of  $^{13}\text{C}_6$  vanillin.

234 The quantification was carried out using a response factor in order to avoid the instrumental signal  
235 fluctuations.

236 Trueness is an important parameter to evaluate during method validation. It refers to the degree of  
237 closeness of the determined value to the known “true” value. It is expressed as a percent error,  
238 calculated as  $(Q-T)/T \times 100$  where Q is the determined value and T is the “true value”. The error for cis-  
239 pinonic acid and pinic acid was calculated performing the same pre-analytical procedure achieved with  
240 the environmental samples. For the evaluation of the extraction yield to estimate the procedural  
241 extraction efficiency the isotopically labelled  $^{13}\text{C}_6$  vanillin was added after the PTFE filtration. Table 1  
242 shows the validation values just described for cis-pinonic acid and pinic acid. The internal standard  
243 method provided an error percentage and CV%  $< \pm 10\%$  for each compound. The recovery of the  
244 analytical procedure for the investigated acids ranged between  $66 \pm 7$  and  $85 \pm 5\%$ ; Parshintsev et al.  
245 (Parshintsev et al., 2010) reached extraction levels from  $77 \pm 9$  to  $96 \pm 4\%$ : the values are close with the  
246 extraction levels of QFF, but they are higher than aluminium plates. This means that the quartz fiber  
247 filter is a better substrate for extraction. In Table 1 is reported the mean absolute blank amount which  
248 was subtracted from the analytical results. The MDL and The MQL were evaluated through 5  
249 procedural blanks, i.e. 5 aluminium substrates and 5 quartz substrates in which it has been added only  
250 the  $^{13}\text{C}_6$  vanillin after the extraction. The MDL and the MQL were evaluated as 3 and 10 times the  
251 standard deviation of these field blanks. Even though the fiber filter has a higher MDL compared to the  
252 aluminium plate due its porosity, the values are quite similar to the procedural blanks, demonstrating a  
253 minimal contamination during the operation before and after the sampling.

#### 254 255 **Cis-pinonic and pinic acid in the urban atmospheric aerosol**

256 Cis pinonic and pinic acids were determined in the atmospheric aerosol collected in the urban area of  
257 Mestre-Venice (Italy) from 14 March to 13 May 2017. The total concentration of each acid, calculated  
258 as the sum of their size distributions in all aerosol samples, has an average value of  $0.3 \text{ ng m}^{-3}$ . Cis-  
259 pinonic acid was usually found in the ultrafine fraction ( $< 0.056 \text{ nm}$ ), and its concentrations ranged from  
260 below MDL to  $0.8 \text{ ng m}^{-3}$ . Instead pinic acid concentrations ranged from below MDL to  $0.6 \text{ ng m}^{-3}$ . The  
261 concentration values found in this study are lower of an order of magnitude than the investigation  
262 reported in literature (Fu et al., 2009; Kavouras et al., 1998; Kavouras and Stephanou, 2002; Sheesley

263 and Kenski, 2004; Yu et al., 1999; Zhang et al., 2010) (Table 2). These authors collected the samples  
264 very close to conifers and deciduous areas, while in the present study the sampling site was just near a  
265 restrict area of deciduous trees. To ensure that is so, Lamb et al. (Lamb et al., 1987) demonstrated how  
266 monoterpenes are mostly formed near coniferous and deciduous trees, although the formation of the  
267 photo-oxidation products follows a partially unknown mechanism, forming an abundance of SOA  
268 compounds with a wide range particles diameter. Consequently, fine and ultra-fine particles can be  
269 transferred for long distances, according to the atmospheric conditions. The location of aerosol site  
270 collection certainly influences the concentrations of photo-oxidation products of  $\alpha$ -pinene, although  
271 other parameters can have an important role. The sampling site provide even the chemical atmospheric  
272 evolution of the two terpenoic acids. In fact, the features of the urban area, considering even the huge  
273 distance from coniferous sources, is such as to give information about the atmospheric transformations.  
274 In literature it is demonstrated that the highest concentrations for cis-pinonic acid and pinic acid are  
275 measured in spring and summer months (Sheesley and Kenski, 2004; Zhang et al., 2010). The results of  
276 this study agree with these observations (Figure 2) for cis-pinonic acid, because it increased its  
277 concentrations form March to May, while pinic acid has a different behaviour. The major  
278 concentrations of cis-pinonic acid have been found in the three samples of April 26-29 and May 3-6 and  
279 May 10-13, while pinic acid is most concentrated in the collecting periods of March 14-18, 18-22 and  
280 May 3-6.

281 It is known how cis-pinonic acid is a high/semi-volatile compound (Zhang et al., 2010) and it was  
282 detected in the gas and particulate phase in forests atmosphere (Kavouras et al., 1999, 1998; Kavouras  
283 and Stephanou, 2002; Pio et al., 2001). Figure 2 and the box-plot diagram of Figure 3 show that cis-  
284 pinonic acid was only found in the ultrafine particles (mostly distributed below 0.056 nm diameter) and  
285 this suggests that it is a typical first-generation reaction product (Jimenez et al., 2009) with a gas-to-  
286 particle process (Anttila et al., 2005; Pio et al., 2006), due to its abundant presence in the gas phase.

287 Several studies explained that pinic acid derives from pinonic acid through photo-oxidation processes  
288 (Lamb et al., 1987; Larsen et al., 1999; Librando and Tringali, 2005; Noziere et al., 1999). The  
289 scientific literature describes pinic acid as low/semi-volatile compounds that is mainly present in the  
290 submicrometer fraction of ambient aerosols (Alves et al., 2000; Pio et al., 2006). However, considering

291 that SOAs from monoterpene photo-oxidation constitute good CCN (cloud condensation nuclei) in the  
292 atmosphere (Huff Hartz et al., 2005; O'Dowd et al., 2002), cis-pinonic acid can undergo the photo-  
293 oxidation in particulate phase creating pinic acid that, being also a good CCN, grows forming fine  
294 aerosol with a greater diameter, up to 1  $\mu\text{m}$  (Spurny, 2000). Coarse particles (diameter  $> 1 \mu\text{m}$ ) are  
295 mostly emitted to the atmosphere during mechanical processes from both natural and anthropogenic  
296 sources. A further explanation of the presence of pinic acid in coarse particles might be the result of  
297 condensation of pinic acid, produced by the gas-phase photo-oxidation of cis-pinonic acid, onto larger  
298 existing aerosol particles or of particle coagulation, especially during long-range transport (Herckes et  
299 al., 2006; Wang et al., 2009; Zangrando et al., 2016, 2013). Very often the equilibrium-phase  
300 partitioning depends strongly on temperature, because of the variation of vapour pressures of the  
301 condensing compounds with temperature (Seinfeld and Pankow, 2003). Figure 3 show that the trend  
302 concentration, fractions distribution and the formation processes of pinic acid tends to distribute in all  
303 the fractions. It can be explained with a tri-modal view: nucleation (Aitken) mode (particle diameter  $<$   
304  $0.1 \mu\text{m}$ ), accumulation mode (particle diameter:  $0.1 \mu\text{m} > d > 1 \mu\text{m}$ ) and coarse mode (particle diameter  
305  $> 1 \mu\text{m}$ ).

306 In this study cis-pinonic acid and pinic acid don't have a clear correlation with temperature (Figure S-  
307 3), probably due to the different atmospheric conditions undergo by the particles during the transport  
308 processes. Moreover, there is not a clear relationship among the trend concentrations of the acids and  
309 precipitations, relative humidity and wind directions and intensity (relative humidity, precipitation and  
310 wind rose are shown in the supporting material, Figure S-2 and S-3). Pio et al. (Pio et al., 2006) has  
311 shown, using a cascade impactor, how pinonic and pinic acids appeared mainly in fine particles ( $<0.69$   
312  $\mu\text{m}$  diameter): the distribution differences with this study is due to the distance from the sources. To our  
313 knowledge this is the first study in which we can observe the behaviour of cis pinonic acid and pinic  
314 acid in aged aerosol in an urban environment.

315 **Conclusions**

316 In this study a method for the quantification in particulate matter of two terpenic acid, cis-pinonic acid  
317 and pinic acid, using a HPLC-ESI(-)-MS/MS system was developed. We obtained a sensitive method  
318 with instrumental detection limits of 1.6 and 1.2 ngL<sup>-1</sup>, respectively. To our knowledge, this is the most  
319 sensitive method to quantify these target acids. The analytical procedure was validated to accurately  
320 quantify these compounds in the aerosol samples through the estimation of trueness, repeatability, and  
321 recovery.

322 The HPLC-MS/MS method developed in this study was applied to the atmospheric samples collected in  
323 Mestre-Venice, to characterise particle size distribution of cis-pinonic acid and pinic acid. The sampling  
324 was conducted using a MOUDI II cascade impactor to discriminate the particle size from 18 µm to <  
325 0.056 µm. This is the first study about the characterization of the pinic and cis-pinonic in the ultrafine  
326 particles. During the spring 2017 fourteen different samples demonstrated that cis-pinonic acid is  
327 mostly distributed in the ultra-fine fraction (below 0.056 µm diameter) while pinic acid show a steady  
328 distribution among the 12 fractions. Both acids seem don't have a clear correlation with temperature.

329

330 **Acknowledgements**

331 The research was supported by the National Research Council of Italy (Consiglio Nazionale delle  
332 Ricerche, CNR). The research leading to these results has received funding from the European Research  
333 Council under the European Union's Seventh Framework Programme (FP7/2007-2013) / ERC Grant  
334 agreement n° 267696 – EARLYhumanIMPACT. The authors also acknowledge Elga Lab Water (High  
335 Wycombe, UK) for providing the ultrapure water systems. Data and information on local meteorology  
336 were obtained from FISTEC-Mestre (IUAV, Venice - Environmental engineering physics laboratory,  
337 website: [fistec.iuav.it](http://fistec.iuav.it)). We would also like to thank Dr. Natalie M. Kehrwald for the revision of our  
338 manuscript.

339



340 **Table 1**

Compound	Error %	Recovery %	CV%	Blank (ng)	MDL (ng)	MQL (ng)
Aluminium						
Cis-pinonic acid	5.9	77±1	2	2.3±0.2	0.7	2.5
Pinic acid	-9.2	66±7	10	3.1±0.4	1.2	4.1
Quartz QFF						
Cis-pinonic acid	2.3	80±2	2	3.0±0.8	2.3	7.5
Pinic acid	0.9	85±5	6	2.7±0.6	1.9	6.2

341 Table 1. Average errors (%), recovery (%), CV%, blank (ng), MDL and MQL (ng)

342 **Table 2**

Compound	Average conc. ng m <sup>-3</sup>	Location, time
Cis-pinonic acid	0.3 ± 0.3	Mestre-Venice, Italy, March-May 2016 (this study)
	1.22 ± 1.33	Mainz, Germany, May 2006–June 2007 (Zhang et al., 2010)
	0.069 ± 0.023	Canadian Arctic, February–June 1991 (Fu et al., 2009)
	18 ± 31	SMEARII station, Finland, August 2007 (Parshintsev et al., 2010)
	40.5 ± 67.5	Alabama, USA, May 2004–April 2005 (Sheesley and Kenski, 2004)
	9.7 ± 11	Pertouli, Greece, August 1998 (Kavouras and Stephanou, 2002)
Pinic acid	0.3 ± 0.2	Mestre-Venice, Italy, March-May 2016 (this study)
	2.32 ± 2.72	Mainz, Germany, May 2006–June 2007 (Zhang et al., 2010)
	0.51 ± 0.40	Canadian Arctic, February–June 1991 (Fu et al., 2009)
	1 ± 9	SMEARII station, Finland, August 2007 (Parshintsev et al., 2010)
	0.54 ± n.d.	Nova Scotia, Canada, July 1996 (Yu et al., 1999)
	0.5 ± n.d.	San Bernadino, Canada, September 1998 (Yu et al., 1999)
	2.4 ± 1.5	Pertouli, Greece, August 1998 (Kavouras and Stephanou, 2002)

343 Table 2. Average TSP concentration comparison with other studies of investigated organic compounds.

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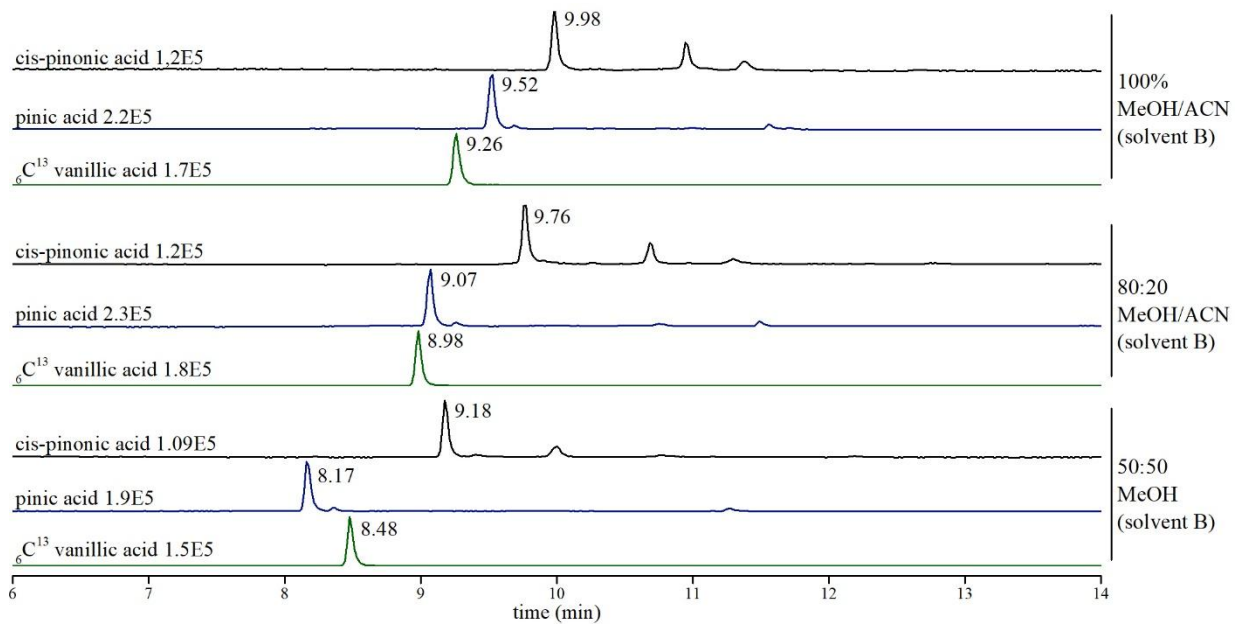
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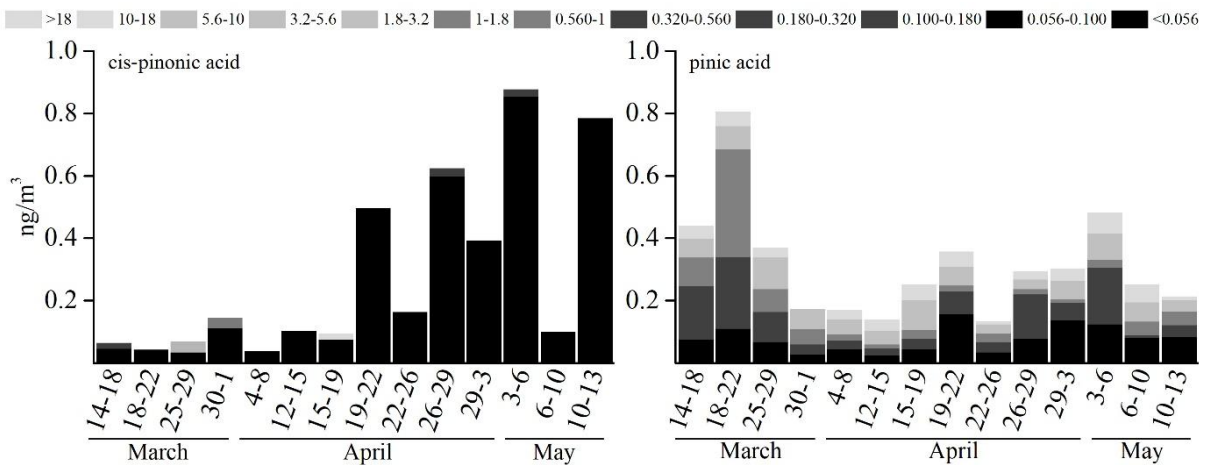
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351 **Figure 1**



352  
 353 Figure 1. Chromatograms of the target acid and their internal standard tested with 3 different mobile  
 354 phases from 6 to 14 minutes. Each ion chromatogram is related to most intense ions of MRM method  
 355 for a standard solution of the acids with an average concentration of  $50 \mu\text{g L}^{-1}$  (cis-pinonic acid  
 356  $183.2/139.0$ ; pinic acid  $185.2/140.9$ ;  $^{13}\text{C}^{13}$  vanillic acid  $156.9/141.8$ ).

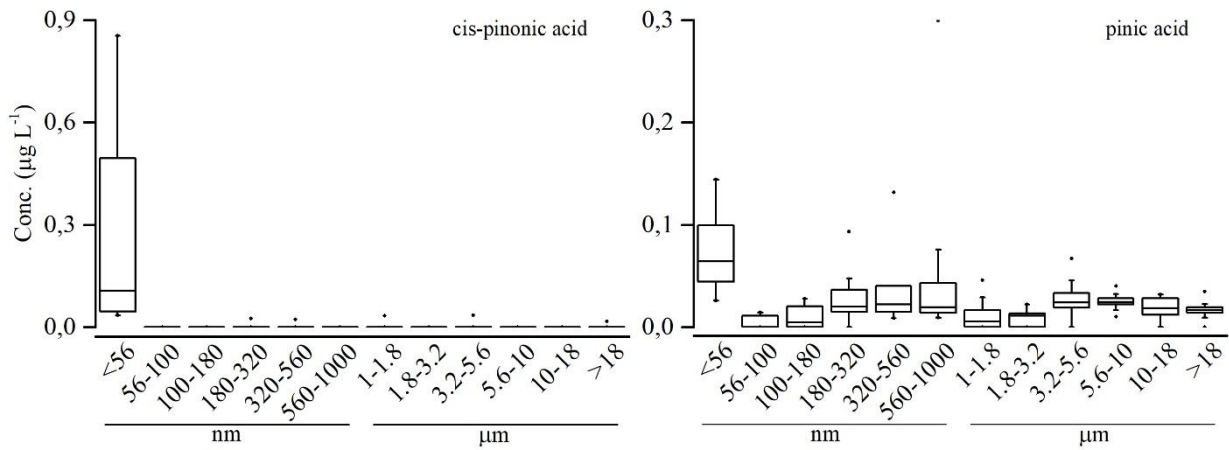
357 **Figure 2**



358  
 359 Figure 2. Monthly variations and fractions distribution of target acids ( $\text{ng m}^{-3}$ ).

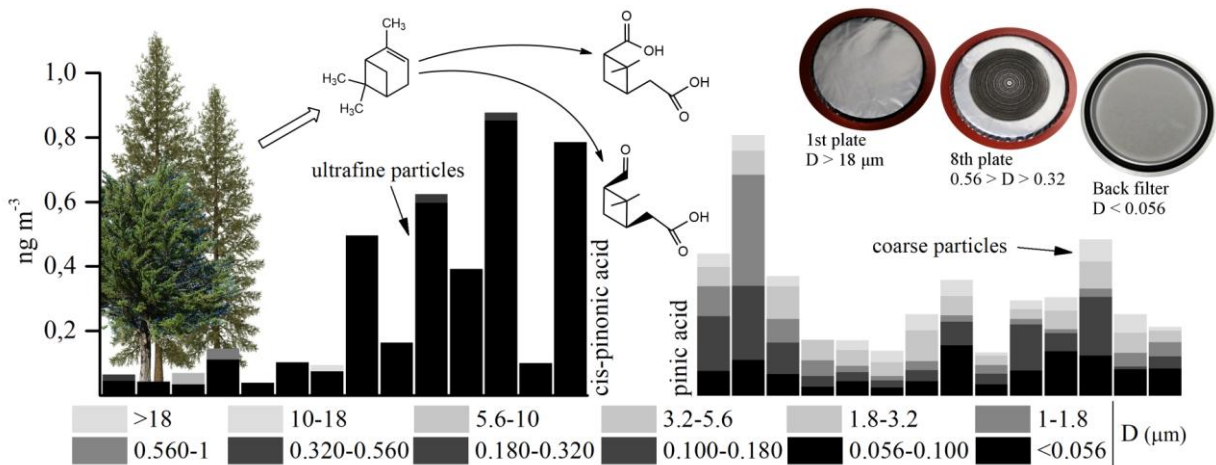
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364 **Figure 3**



365  
 366 Figure 3. Box-plot diagram of cis-pinonic acid and pinic acid according to the particles diameter. The  
 367 line inside the box is referred to the median.  
 368

369 **FOR TOC ONLY**



370  
 371

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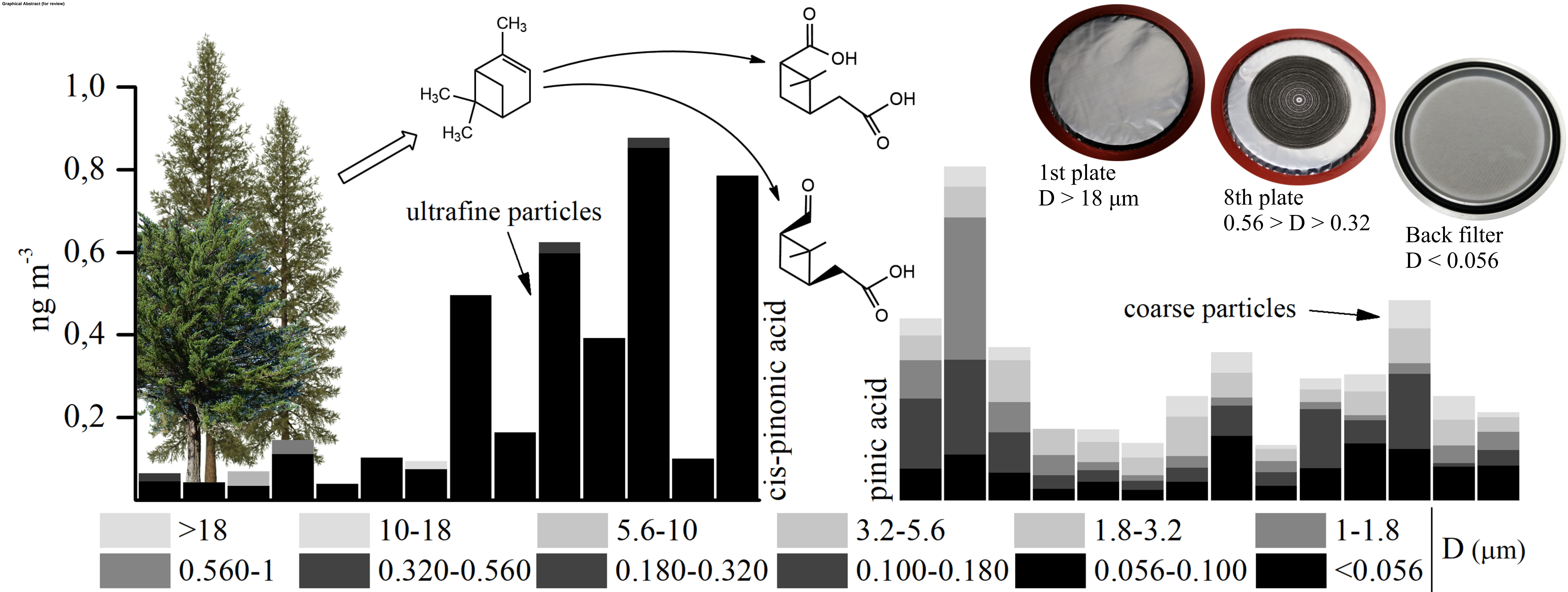
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499







## **Highlights**

A sensitive HPLC-MS/MS method was developed to quantify cis-pinonic and pinic acids.

Aerosol particle size distribution of cis-pinonic and pinic acids was evaluated.

For the first time cis-pinonic and pinic acids were found in the nanoparticles.

Figure 1  
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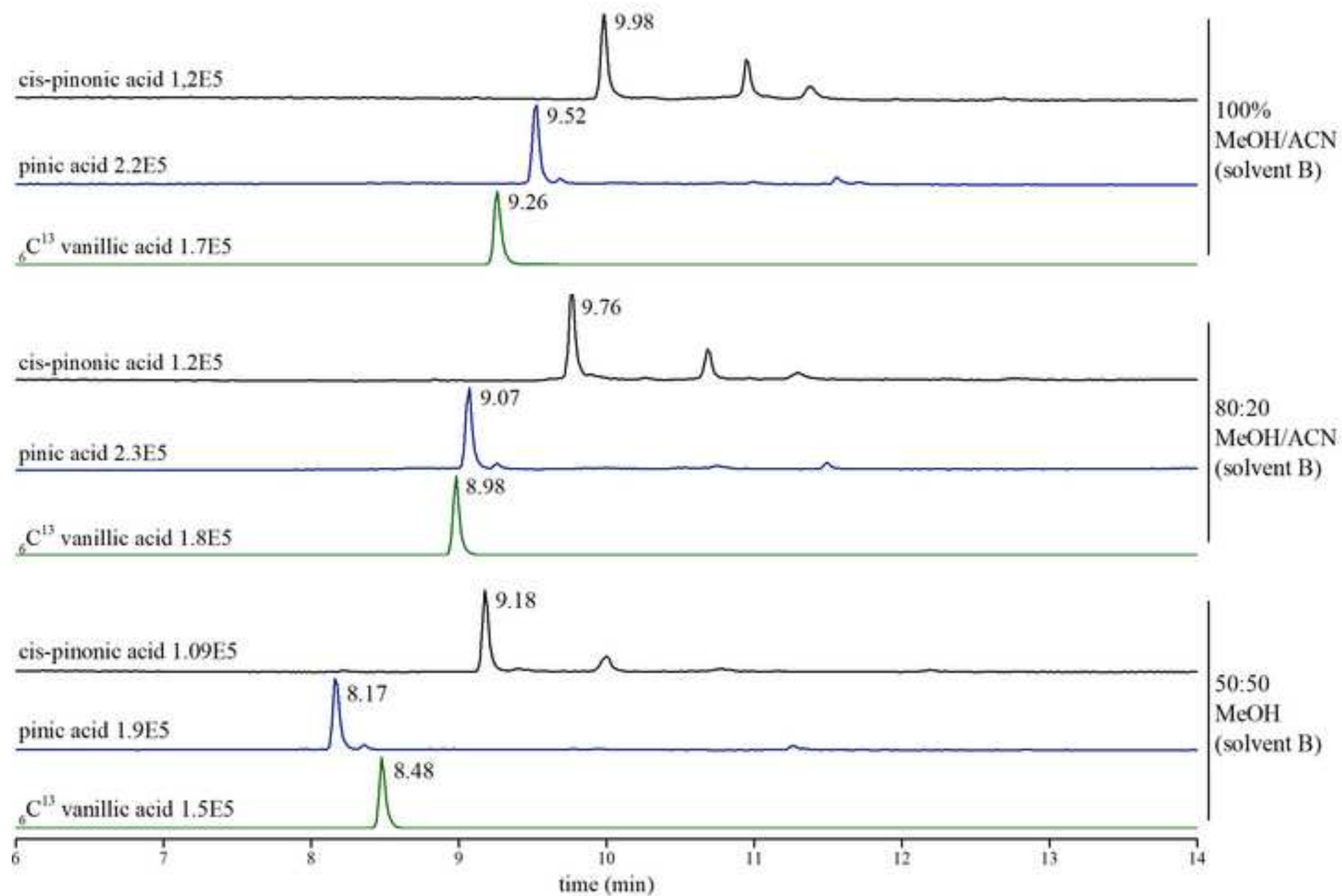


Figure 2  
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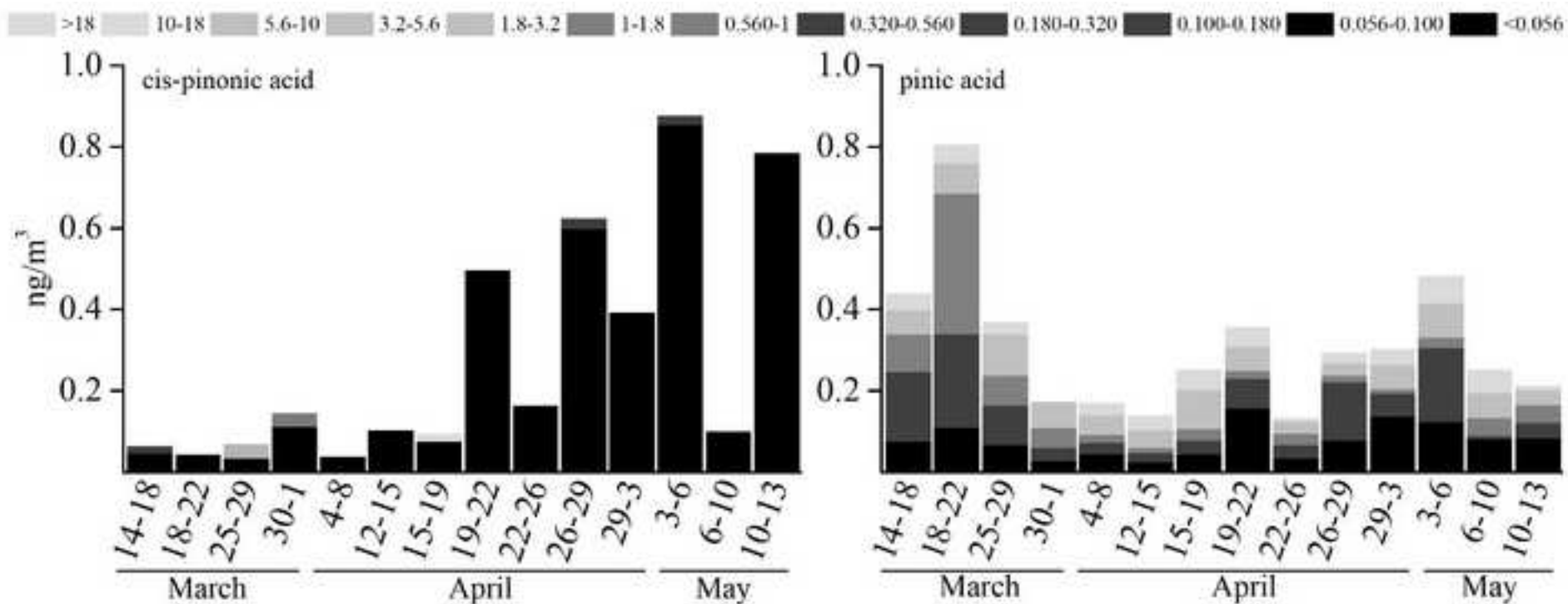
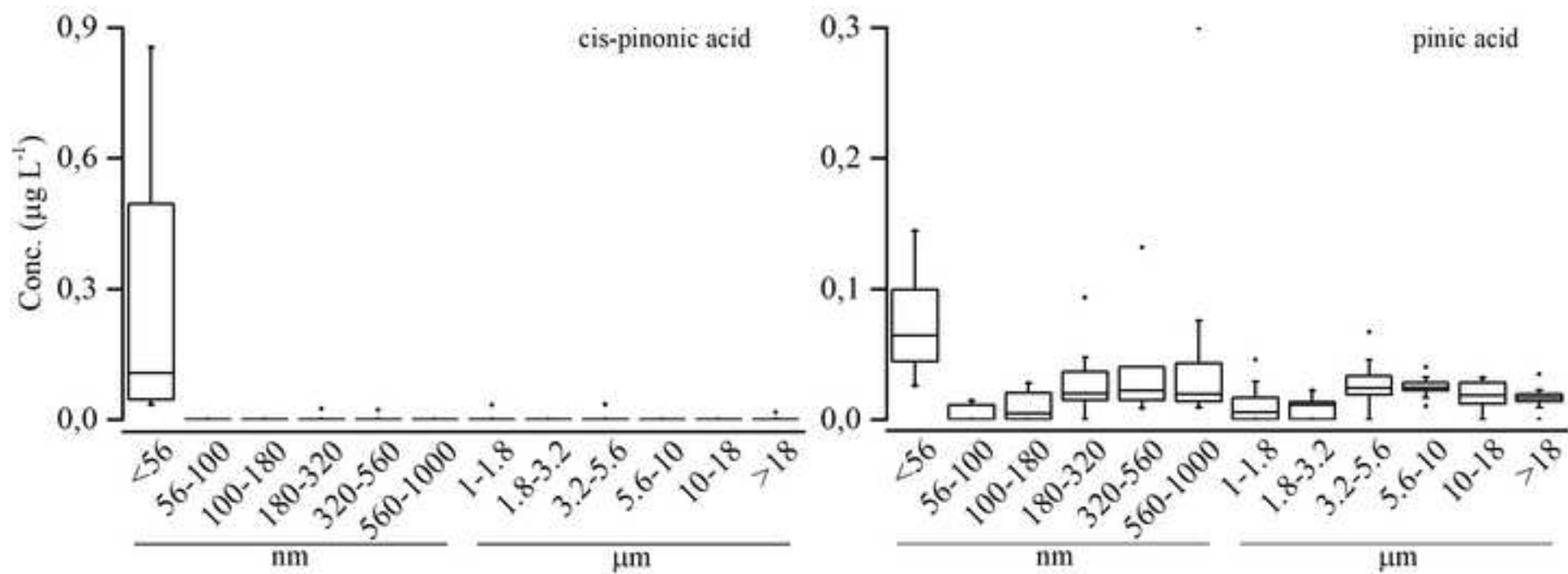


Figure 3  
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**Table 1**[Click here to download Table: Table 1.docx](#)**Table 1**

Compound	Error %	Recovery %	CV%	Blank (ng)	MDL (ng)	MQL (ng)
Aluminium						
Cis-pinonic acid	5.9	77±1	2	2.3±0.2	0.7	2.5
Pinic acid	-9.2	66±7	10	3.1±0.4	1.2	4.1
Quartz QFF						
Cis-pinonic acid	2.3	80±2	2	3.0±0.8	2.3	7.5
Pinic acid	0.9	85±5	6	2.7±0.6	1.9	6.2

Table 1. Average errors (%), recovery (%), CV%, blank (ng), MDL and MQL (ng)

**Table 2**[Click here to download Table: Table 2.docx](#)**Table 2**

Compound	Average conc. ng m <sup>-3</sup>	Location, time
Cis-pinonic acid	0.3 ± 0.3	Mestre-Venice, Italy, March-May 2016 (this study)
	1.22 ± 1.33	Mainz, Germany, May 2006–June 2007 (Zhang et al., 2010)
	0.069 ± 0.023	Canadian Arctic, February–June 1991 (Fu et al., 2009)
	18 ± 31	SMEARII station, Finland, August 2007 (Parshintsev et al., 2010)
	40.5 ± 67.5	Alabama, USA, May 2004–April 2005 (Sheesley and Kenski, 2004)
	9.7 ± 11	Pertouli, Greece, August 1998 (Kavouras and Stephanou, 2002)
Pinic acid	0.3 ± 0.2	Mestre-Venice, Italy, March-May 2016 (this study)
	2.32 ± 2.72	Mainz, Germany, May 2006–June 2007 (Zhang et al., 2010)
	0.51 ± 0.40	Canadian Arctic, February–June 1991 (Fu et al., 2009)
	1 ± 9	SMEARII station, Finland, August 2007 (Parshintsev et al., 2010)
	0.54 ± n.d.	Nova Scotia, Canada, July 1996 (Yu et al., 1999)
	0.5 ± n.d.	San Bernadino, Canada, September 1998 (Yu et al., 1999)
	2.4 ± 1.5	Pertouli, Greece, August 1998 (Kavouras and Stephanou, 2002)

Table 2. Average TSP concentration comparison with other studies of investigated organic compounds.

**Supplementary Material**

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